Elsevier Editorial System(tm) for Science of

the Total Environment

Manuscript Draft

Manuscript Number: STOTEN-D-20-01110R2

Title: Comment on: "A novel approach to peatlands as archives of total cumulative spatial pollution loads from atmospheric deposition of airborne elements complementary to EMEP data: priority pollutants (Pb, Cd, Hg)" by Ewa Miszczak, Sebastian Stefaniak, Adam Michczyński, Eiliv Steinnes and Irena Twardowska.

Article Type: Letter to the Editor

Keywords: Lead, immobility, ombrotrophic peatland, bog, geochemistry, enrichment factor, metal accumulation rate

Corresponding Author: Dr. Francois De Vleeschouwer,

Corresponding Author's Institution: CNRS

First Author: Francois De Vleeschouwer

Order of Authors: Francois De Vleeschouwer; Sandrine Baron; Joanna M. CLoy; Maxime Enrico; Vojtech Ettler; Nathalie Fagel; Heike Kempter; Malin Kylander; Chuxian Li; Jack Longman; Antonio Martinez-Cortizas; Samuel Marx; Nadine Mattielli; Timothy Mighall; Tiina M. Nieminen; Natalia Piotrowska; Xabier Pontevedra Pombal; Steve Pratte; Virginie Renson; William Shotyk; Emma Shuttleworth; Jaroslaw Sikorski; Nicola Stromsoe; Julie Talbot; Clemens von Scheffer; Dominik Weiss; Claudio Zaccone; Gaël Le Roux

Manuscript Region of Origin: ARGENTINA

UMI-3351 INSTITUT FRANCO-ARGENTIN D'ETUDES SUR LE CLIMAT ET SES IMPACT INSTITUTO FRANCO-ARGENTINO DEL ESTUDIO DEL CLIMA Y SUS IMPACTOS

Subject: Critical Comment on Miszczak et al (2020) – R2

Buenos Aires, April 6th, 2020

To Science of the Total Environment Co-Editors in chief Damià Barceló and Jay Gan.

Dear Dr Barcelo and Dr Gan,

We hereby submit our revised version (R2) of our letter to Editor to comment the paper that Miszczak et al. (2020) recently published in Science of the Total Environment [\(https://doi.org/10.1016/j.scitotenv.2019.135776\)](https://doi.org/10.1016/j.scitotenv.2019.135776).

We reiterate the importance of the mistakes those authors have made, together with ignoring the vast majority of literature in the field. This paper should not have passed the review process and while reviewer 1 claims our tone is insulting and attacking, we really don't think so. If any, it is quite the opposite as our community has felt attacked and insulted, otherwise we would not be 27 to co-sign this comment. As such, this paper is neither a research group argument nor a discussion. It is just a community putting straight back the many major flaws of this paper, for the sake of scientific accuracy and for the STOTEN readers.

We kept out short Supplementary Material that gives the opportunity to list the entire literature as we think it is crucial to show how many papers are contesting what Miszczak et al. (2020) have claimed and that our community finds it imperative to correct their many scientifically invalid claims. As it contains other corrections and criticisms, we do not think it could be appropriately converted in a "data in brief format", but are happy to further discuss that point with you.

We reaffirm that the approach that the peat geochemistry community follows to reconstruct past anthropogenic activities using lead profiles in ombrotrophic peat has been the pillar of most of the papers dealing with Pb as an indicator of global pollution over the world as you can see in the table in the Supplementary Material.

We insist on the fact that the vast majority of our community was appalled by the publication of Miszczak et al. (2020) -as the reviewers point out this paper should not have passed the revision- and have therefore co-signed this Letter to Editor. We find it fundamental to publish it. We therefore hope you are satisfied with this version of our comment that we hope to be final.

UMI-3351 INSTITUT FRANCO-ARGENTIN D'ETUDES SUR LE CLIMAT ET SES IMPACT INSTITUTO FRANCO-ARGENTINO DEL ESTUDIO DEL CLIMA Y SUS IMPACTOS

On the behalf of: Baron S. (CNRS, Toulouse, France), Cloy J.M. (Scotland's Rural College, Edinburgh, UK), Enrico M. (Harvard, USA), Ettler V. (Charles University Prague, Czech Republic), Fagel N. (U. Liège, Belgium), Kempter H., Kylander M. (U. Stockholm, Sweden), Li C. (GET, Toulouse, France), Longman J. (U. Oxford, UK), Martinez-Cortizas A. (U. Santiago de Compostela, Spain), Marx S. (U. Wollongong, Australia), Mattielli N. (U. Brussels, Belgium), Mighall T. (U. Aberdeen, UK), Nieminen T.M. (Natural Resources Institute, Helsinki, Finland), Piotrowska N. (GADAM Center, Poland), Pontevedra Pombal X. (U. Santiago de Compostela, Spain), Pratte S. (Zhejiang University), Renson V. (U. Missouri, USA), Shotyk W. (U. Alberta, Canada), Shuttleworth E. (U. Manchester, UK), Sikorski J. (GADAM Center, Poland), Talbot J. (U. Montreal, Canada), von Scheffer C. (U. Kiel, Germany), Weiss D. (Imperial College London, UK), Zaccone C. (U. Verona, Italy), Le Roux G. (CNRS, Toulouse, France)

Sincerely Yours,

François De Vleeschouwer CNRS Researcher, Adjunct Director Instituto Franco-Argentino para el Estudio del Clima y sus Impactos UMI IFAECI, CNRS-CONICET-UBA-IRD, Buenos Aires, Argentina Tel: +54 11 5285-8458, fdevleeschouwer@cima.fcen.ub

Reviewer #1: The authors have corrected the errors that I noted and have re-structured the paper so that it is improved substantially. The statements that are made are now factually correct. The remaining concern to me is the "tone" of the paper. While I appreciate that the authors' tone is less severe than in the previous version, at least in the abstract, the conclusion changes and once again it reads like an attack on the Miszazak et al paper. I strongly contend that if a paper is published primarily as a comment on another paper - the authors of the original paper should be allowed to comment. Is, as seems the case, the authors appear to be writing a paper that reiterates a scientific point that others are misinterpreting then I still feel it could be toned down without compromising this paper if the primary intent is to confirm that Pb is largely immobile in ombrotrophic peat and can reconstruct atmospheric Pb deposition. If it is the intent to criticize the Miszazak et al paper then it must go to these authors.

We thank the reviewer for his/her comment.

Let's not reverse things here. If anybody has been attacked our insulted, it is our community (hence we are so many to sign up this comment). It's not the opposite. This paper is appalling and should not have been published. While we can understand that STOTEN would not take the extreme decision of removing such flawed paper, we are left with the choice of putting things back. Hence there is indeed not much discussion. We merely recall the facts.

We have once again attempted to change the tone of our manuscript, although again, we were merely putting the facts, and strongly believed that yes, Miszczak et al failed to understand crucial concepts. There is no other explanation as the reviewer think there could be.

We however contest that the Miszczak et al. paper is deeply flawed and ignores a vast body of literature and for that deserves criticism. However, we agree it is important to ensure we focus on how Miszczak et al. interpret the data, but nevertheless think it is important to point out their key oversights.

About the comment: There seem to be a misunderstanding here. As we had already pointed out in our previous revised version, yes this is a comment. And yes, Miszczak et al. will have the right to reply to our comment, but once and only once our comment has passed the regular review process and is accepted for publication. Once our comment is accepted, STOTEN will made Miszczak et al. aware of it and allow them to reply. The STOTEN editor already confirmed us this process: Miszczak et al. will have the opportunity to reply. But of course they do not review our manuscript before publication.

My suggestions include:

1. Perhaps changing the title? If it is a comment on another paper those authors should be allowed to comment. As it in the review stage it will not compromise the publication.

We have not changed our title as it is indeed a comment, and Miszczak et al. will have the right to reply. We hope this is now clear with the explanations provided above. Such a title can be accepted (see past publications in STOTEN) and will appear in the "letter to the editor" section of a STOTEN volume, together with the reply of Miszczak et al.

We feel it is important to retain a direct reference to the Miszazak et al.'s paper as our contribution deals directly with the misapprehensions in that work.

2. The first sentence/abstract could simply be "Recent work has suggested that … In this paper we reiterate the evidence collected over 40 years of study that demonstrates Pb is largely immobile in ombrotrophic peat cores and highlight some errors that may lead to misinterpretation of this valuable archival record"

Changed as suggested.

3. The paper could be re-written slightly so that the points are retained but is not a direct criticism of this paper - for example.

Well, this is a direct criticism of that paper which is full of major flaws and that should not have been published… We have anyways adjusted the text.

4. Ombrotrophic peatlands are well-established archives of past atmospheric deposition of trace elements. After more than 40 years of investigation (see Table SI1), there is a consensus that Pb is largely immobile in ombrotrophic peatlands. We raise a series of critical issues regarding Miszczak et al. (2019) to clarify how ombrotrophic peatlands have allowed the reconstruction of past patterns of atmospheric metal contamination in the environment. We do not aim to provide a complete review of Pb in mires, however because the trophic status of a peatland is crucial to studies of past atmospheric metal deposition (see Supplementary Material for further details), we refer to a selection of studies relevant to the type of peatland that Miszczak et al. (2019) used, namely ombrotrophic peatlands, also known as bogs.

Could be written:

Ombrotrophic peatlands are well-established archives of past atmospheric deposition of trace elements. After more than 40 years of investigation (see Table SI1), there is a consensus that Pb is largely immobile in ombrotrophic peatlands. To avoid some confusion raised by a recent publication (Miszazak et al. 2019), we seek to clarify how ombrotrophic peatlands have allowed the reconstruction of past patterns of atmospheric metal contamination in the environment. We do not aim to provide a complete review of Pb in mires, however because the trophic status of a peatland is crucial to studies of past atmospheric metal deposition (see Supplementary Material for further details), we refer to a selection of studies relevant to ombrotrophic peatlands, also known as bogs.

We have re-phrased this section, although we don't use the Reviewer's exact wording, but have modified this section in a similar way.

5. I do not wish to go through the entire paper this way but hopefully my point is taken. The authors can still make all their valid points without continuously commenting on the Miszczak paper, but still make it clear that they had concerns with this paper. I would have no issues if the paper was written in this style. If it remains as presented, as a scientist myself, if someone where to directly attack my work I would want the chance to respond and I contend that it must go to the authors of the paper in question if that is the path chosen. For example, reading the sentence on line 106 is directly insulting the authors and makes an assumption that they "were unaware". Perhaps they are aware but had a different reason for doing what they did? On line 109, the authors claim "they failed to comprehend" - this is derogatory terminology. My suggestions are easy to adopt: Section 2 is fine if it stops at line 47. In section 3, just delete the sentence beginning on line 60.

We have amended the manuscript as suggested throughout

6. Most changes are needed from sections 3.1 onwards, but they are simply editorial and are needed to change the tone and can easily be done, without compromising this paper in any way [I could likely do it in 30 minutes but it is not my paper]. I leave it to the authors to decide if they wish to do this but I do not recommend the paper be published unless they do. I do not wish to see reputable journals become a venue for arguments between research groups.

We take the Reviewers point and have amended our manuscript in light of these comments. This has necessitated some re-organisation of the manuscript to re-focus our comment on more explicitly on the utility of bogs and the necessary steps needed to reconstruct Pb contamination histories.

We feel it is important to also point out that we are not a research group, but are a group of authors representative of number of research groups working within this field. While a number of the authors have previously worked together, others have not. Nevertheless, when this paper appeared, we were all outraged by the low quality of the paper, the mis-interpretation of previous work (as well as some attacks), and yes, felt Miszczak et al. were insulting our community. We therefore felt the urge to write this comment. We however understand the reviewer as to that we should not reply by attacking or insulting, and have adjusted our manuscript once more, hopefully to the liking of the reviewer and editor.

A novel approach to peatlands as archives of total cumulative spatial pollution loads from atmospheric deposition of airborne elements complementary to EMEP data: priority pollutants (Pb, Cd, Hg)".

by Ewa Miszczak, Sebastian Stefaniak, Adam Michczyński, Eiliv Steinnes and Irena Twardowska.

Authors

De Vleeschouwer F.^{1*}, Baron S.², Cloy J.M.³, Enrico M.⁴, Ettler V.⁵, Fagel N.⁶, Kempter H.⁷, Kylander M.⁸, Li C.⁹, Longman J.¹⁰, Martinez-Cortizas A.¹¹, Marx S.¹², Mattielli N.¹³, Mighall T^{14} , Nieminen T.M.¹⁵, Piotrowska N.¹⁶, Pontevedra Pombal X.¹¹, Pratte S.¹⁷, Renson V.¹⁸, Shotyk W.¹⁹, Shuttleworth E.²⁰, Sikorski J.¹⁶, Stromsoe N.²¹, Talbot J.²², von Scheffer C.²³, D. Weiss 24,25 , Zaccone C. 25 , Le Roux G. 26

- 1. Instituto Franco-Argentino para el Estudio del Clima y sus Impactos (UMI IFAECI/CNRS-CONICET-UBA-IRD)
- 2. Laboratoire TRACES (CNRS, Université de Toulouse), France
- 3. Scotland's Rural College, Edinburgh, UK
- 4. Harvard John A. Paulson School of Engineering and Applied Science, Harvard University, Cambridge, MA, USA
- 5. Charles University, Prague, Czech Republic
- 6. AGEs, Département de Géologie, Université de Liège, Belgium
- 7. Welzheimer Str. 14, D-71566 Althuette, Germany
- 8. Department of Geological Sciences and the Bolin Centre for Climate Research, Stockholm University, Sweden
- 9. Geoscience Environnement Toulouse (CNRS-UPS-IRD-CNAP-CNES), France
- 10. School of Geography and the Environment, University of Oxford, UK
- 11. Departamento de Edafoloxia e Quimica AgricolaEcoPast (GI-1553), Universidad de Santiago de Compostela, Spain
- 12. GeoQuEST Research Centre, School of Earth, Atmospheric and Life Sciences, University of Wollongong, Australia
- 13. Laboratoire G-Time, Université Libre de Bruxelles, Belgium
- 14. School of Geosciences, University of Aberdeen, UK
- 15. Natural Resources Institute Finland Luke, Helsinki, Finland
- 16. Silesian University of Technology, Institute of Physics-CSE, GADAM Center, Gliwice, Poland
- 17. Department of Geography, School of Earth Sciences, Zhejiang University, Hangzhou, China

- 18. Research Reactor, University of Missouri, USA.
- 19. Department of Renewable Resources, University of Alberta, Edmonton, Canada
- 20. Department of Geography, University of Manchester, UK
- 21. College of Engineering, IT and Environment, Charles Darwin University, Australia
- 22. Département de Géographie, Université de Montréal, Canada
- 23. Institute for Ecosystem Research, Kiel University, Germany
- 24. Imperial College London, UK
- 25. 25 Princeton University, Princeton, USA
- 26. Department of Biotechnology, University of Verona, Italy
- 27. EcoLab, Université de Toulouse, CNRS, Toulouse, France

* Corresponding author: Dpto. de Ciencias de la Atmosfera y los Oceanos, FCEN, Universidad de Buenos Aires, Intendente Guiraldes 2160 - Ciudad Universitaria, Pabellon II - 2do. Piso, (C1428EGA) Ciudad Autonoma de Buenos Aires – Argentina. [fdevleeschouwer@cima.fcen.uba.ar](mailto: fdevleeschouwer@cima.fcen.uba.ar)

Abstract

A recent paper by Miszczak et al. (20192020); recently presented what they consider to be '*a A novel approach to using peatlands as archives of cumulative atmospheric pollution',* Science of the Total Environment, 705, 135776) examines metal contamination in peat mires in Poland and Norway. The authors conclude that lead (Pb) records in ombrotrophic peatlands cannot be used to reconstruct the chronological history of anthropogenic activities due to post-depositional mobility of the metal. We contest this general conclusion which stands in contrast with a significant body of literature demonstrating that Pb is largely immobile in the vast majority of ombrotrophic peatlands. Our aim is to reaffirm the crucial contribution that peat bog-records have made to our knowledge of atmospheric Pb pollution historiescontamination. Toln addition, we re-iterate the necessity of following accepted protocols to produce reliable records We follow this bywe highlighting what we perceive are eritical issues with the paper published by Miszczak et al. (2019)of anthropogenic Pb contamination in environmental archives.

Keywords

Lead, immobility, ombrotrophic peatland, bog, geochemistry, enrichment factor, metal accumulation rate

1 **1. INTRODUCTION**

2 Ombrotrophic peatlands are well-established archives of past atmospheric deposition of 3 trace elements. After more than 40 years of investigation (see Table SI1), there is a 4 consensus that Pb is largely immobile in ombrotrophic peatlands. We arguecontend that 5 some of the conclusions reached in a recent paper byin Miszczak et al (2020)⁴⁹ are based 6 on misinterpretation or incorrect application of approaches to sample collection ing and data 7 analysis approaches. To avoid such confusion, we seek to clarify here how ombrotrophic 8 peatlands have allowed the reconstruction of past patterns of atmospheric metal 9 contamination in the environment Correcting these misapprehensions is important because 10 **these techniques are well supported by evidence and are essential to producing reliable** 11 records and conclusions.. We raise a series of critical issues regarding Miszczak et al. 12 (2019) to clarify how ombrotrophic peatlands have allowed the reconstruction of past 13 **patterns of atmospheric metal contamination in the environment.** We however do not aim to 14 provide a complete review of Pb in mires, however-because the trophic status of a peatland 15 is crucial to studies of past atmospheric metal deposition (see Supplementary Material for 16 further details), we refer to a selection of studies relevant to ombrotrophic peatlands, also 17 known as bogs. Due to their high atmospheric fidelity, ombrotrophic peatlands have special 18 utility for reconstructing metal contamination records. This was also the type of peatland 19 | investigated by Miszczak et al. (2020).

20 the type of peatland that Miszczak et al. (2019) used, namely ombrotrophic peatlands, also 21 known as bogs.

22

23 2. **THE BEHAVIOUR OF LEAD IN OMBROTROPHIC PEATLANDS**

24 While some studies have suggested that Pb can be mobilized in minerotrophic, riparian, 25 drained or degraded peatlands (e.g. Syrovetnik et al., 2007; Smieja-Król et al., 2010, 2019; 26 Rothwell, 2011; Broder and Biester, 2017) the majority of more than 40 years of literature 27 suggests that Pb is largely immobile in $\frac{1}{100}$ intact-pristine ombrotrophic peat profiles (e.g. De 28 Vleeschouwer et al., 2010a; Marx et al., 2010, Shotyk et al., 2016a,b; Longman et al. 2018; 29 | Fiałkiewicz-Kozieł et al., 2020 and references therein). Although Miszczak et al. (2020) cite 30 literature to support their conclusion of Pb mobility in the bogs they examine, that literature pertains to minerotrophic or disturbed peatlands (e.g. Syrovetnik et al., 2007; Smieja-Król et al., 2010, 2019) and the processes that can promote Pb mobility in those systems are not 33 | applicable to ombrotrophic peatlands.

 The similar elemental and isotopic trends encountered in ombrotrophic peat, lake sediment, ice and herbaria samples (e.g. Rosman et al., 1997; Weiss et al., 1999; Renberg et al., 2001; Farmer et al., 2002; Cloy et al. 2009; Bindler, 2011) and their agreement with anthropogenic emissions patterns (e.g. Shotyk et al., 1998; Mighall et al., 2002; Kylander et al., 2006; De Vleeschouwer et al., 2009a; Marx et al., 2010; Bindler 2011; Cloy et al., 2008, 2009; Allan et al., 2013; Martínez-Cortizas et al., 2016) provide a body of evidence supporting the view that 41 | Pb is largely immobile in ombrotrophic peatlandsbogs. Significantly, stable Pb isotopes in ombrotrophicrecords from peat bogs have consistently been found to accurately reflect temporal variability in source signatures in numerous studies (e.g. in reference *op. cit.*). This would not be the case if post-depositional mobility/isotope mixing was were taking place. 45 Furthermore, in most ombrotrophic peat cores- 210 Pb ages, as determined from the constant 46 Trate of supply (CRS) age-depth models (Appleby and Oldfield, 1978; Appleby, 2001), are- in very good agreement with pollen chronological markers (Appleby et al., 1997), fallout 48 radionuclide chronostratigraphic makers (e.g. from ¹⁴C Bomb Pulse Curve, ¹³⁷Cs and ²⁴¹Am), 49 and tephrochronology in various locations (e.g. Goodsite et al., 2001; Piotrowska et al. 2009; Li et al., 2017; Davies et al., 2018), providing *prima facie* evidence that lead Pb and its isotopes are largely immobile in bogs. Experimental studies lend further support to this (e.g. Vile et al., 1999; Novak et al., 2001). For example, Pb concentrations in the aqueous phase of ombrotrophic peatlands are low (<0.01% of total Pb), while the limited vertical water movement in bogs together with the size of the metal-containing particles in solution limits Pb redistribution (e.g. Shotyk et al., 2016b). Down-washing experiments have also 56 demonstrated that Pb has limited mobility (Hansson et al. 2014, 2015)., while any Tthe 57 | limited mobility that may occurred occur was is not sufficient to compromise the use of Pb to 58 reconstruct pollution historiesy over millennia. We note, however, that the spatial distribution of Pb has tomust be carefully addressed in cases where decomposition and compression integrate signals over longer (decadal and more) timespans (Bindler et al, 2004; Martinez Cortizas et al., 2012). Additionally, Pb behaviour in ombrotrophic peats has been demonstrated to differ from that of mobile elements such as Zn which, in contrast to Pb, displays evidence of vertical diffusion/advection as well as upward plant uptake (e.g. Shotyk 1988; Twardowska et al., 1999; Nieminnen et al., 2002; Weiss et al., 2007).

 In summary, there is a significant body of evidence demonstrating that Pb is largely immobile in bog profiles (see Table S1 and Supplementary Material) that stands in contrast to the **conclusions of Miszczak et al. (2020).**

 Miszczak et al. (2019) fail to recognize the significant body of evidence demonstrating that **Pb is largely immobile in bog profiles (see Table S1 and Supplementary Material). In** addition, the use of it appears they were confused in thinking that literature pertaining to minerotrophic or disturbed peatlands (e.g. Syrovetnik et al., 2007; Smieja-Król et al., 2010, 2019) was to argue for Pb mobility in the bogs they examine is incorrect as the proces 73 which can promote Pb mobility in these systems are not applicable to ombrotrophic 74 peatlands. and thus incorrectly use that evidence to argue for Pb mobility in the bogs they examine. .

3. **HOW TO USE LEAD DATA TO ACCURATELY RECONSTRUCT HISTORICAL CONTAMINATION**

 In the following sections we outline what we consider to be the best practices to ensure accurate reconstruction of atmospheric Pb deposition. We also discuss appropriate 81 approaches to study use Pb contamination pollution records from constructed from peatlands 82 as to records of examine contaminateant sources and to compare with emissions data. The 83 approaches we outline are well established not new and have been described before (e.g. 84 Givelet et al., 2004; De Vleeschouwer et al., 2010b). We hope that these sectionsthis

85 overview corrects any misapprehensions arising from the incorrect application byapproaches 86 used by Miszczak et al. (20192020), do not appear to be aware of theseof commonly applied 87 peat sampling approaches and alsoas well as some misinterpretations previously published 88 data.

89

90 **3.1. Sampling and sub-sampling – Data resolution and geochronology**

91 Correct coring and sub-sampling protocols are important to accurately reconstructing 92 metal contamination deposition chronologies records from peat-mires. The slow accumulation 93 <u>rate of ombrotrophic peatlands means that peat sections oin the order of one vertical</u> 94 centimeter can represent decades of metal accumulation. For example, in European 95 ombrotrophic peatlands, long-term mean peat accumulation rates (i.e. excluding surface 96 vegetation growth) have been estimated to range betweenfrom c. 0.18 to 1 mm yr⁻¹ (e.g. 97 Gorham, 1991; Mäkilä, 1997; Malmer and Wallén, 2004; Pontevedra-Pombal, et al., 2017) 98 depending on the vegetation and climate (e.g. Charman et al., 2013; Pontevedra-Pombal et 99 al., 2019). Because of this, it is common-place for studies to both sample and date the living 100 vegetation at the bog surface (e.g. Farmer et al., 2006; Kempter et al. 2007; Olid et al., 101 2008). This point was illustrated by Givelet et al. (2004) who stated: "*the historical record of* 102 *atmospheric Pb … can depend to a large extent on the methods used to collect, handle, and* 103 *prepare the samples for analysis*". As a result, high-resolution sub-sampling and dating are 104 required to reconstruct decadal-scale atmospheric pollution records. 105 | Here , including because the slow accumulation rate of peat means that peat sections can 106 **represent decades of deposition.** Miszczak et al. (20192020) compare metal ncontaminants 107 $\frac{1}{2}$ in their peat records to European Monitoring and Evaluation Program (EMEP) data (annual 108 trace metal emissions, [https://www.emep.int\)](https://www.emep.int/). Although potentially a very useful undertaking, 109 . Unfortunately, their sampling approach unfortunately greatly reduces the usefulness-utility 110 of such atheir comparison. This is because Miszczak et al. (20192020) used-followed the 111 coring protocol of Steinnes and Sjøbakk (2005) where "*Sphagnum moss … and other plant* 112 *material growing on the surface were removed…before the coring, and the reference surface* *level is thus the interface moss/peat. The thickness of the Sphagnum layer, if present, was always less than 10 cm".* In other words, the authors removed the living/surface vegetation 115 which is an integral part of the ombrotrophic peat deposit, and canpotentially accumulatings **decades of information.**

117

118 **In other words, the authors removed the living/surface vegetation which is an integral part of** 119 **the ombrotrophic peat deposit, and can accumulate decades of information.** Given the slow 120 accumulation rates of bogs, the 40 years of EMEP data are likely, at best, to represent 121 approximately 4 cm of peat accumulation (if surface vegetation is excluded and assuming a 122 | 1mm yr⁻¹ peat accumulation rate). Therefore, the sampling resolution of Miszczak et al. 123 (2020), where peats were subsampled in multi-centimeter increments, combined with their 124 limited use of radionuclide dating, preclude any assessment of recent Pb deposition or 125 comparison with EMEP data from their study. This is especially the case if surface vegetation 126 were removed. ThSlow growth rates, combined with the demonstrated importance of surface 127 vegetation in accumulating metal contaminanttes means that Miszczak et al. (2020) 128 assumption that is approach, i.e., taking the peat/vegetation interface 129 **representingrepresents** the year of coring (in $\frac{1}{100}$ case 1999)₇ is more than likely 130 incorrect. Their approach and would therefore leads to large uncertainties regarding anyin 131 chronology or and any inventory calculations performed thereafter. As previously stated, we 132 consider comparing EMEPPA data with data from ombrotrophic peatlands to be a very 133 worthwhile undertaking, but it requires high-resolution sub-sampling and dating, which is 134 unfortunately not achieved by Miszczak et al. (2020). 135 **The inappropriateness of the Miszczak et al. (2019) sampling approach is demonstrated by** 136 a large number of studies in which the living vegetation has been sampled for Pb and dated

 (e.g. Farmer et al., 2006; Kempter et al. 2007; Olid et al., 2008). This is further illustrated by Givelet et al. (2004) who state: "*the historical record of atmospheric Pb … can depend to a large extent on the methods used to collect, handle, and prepare the samples for analysis*". It 140 is therefore unlikely that Miszczak et al. (2019) can accurately compare multi-centimeter

141 Samples that represent several tens of years with EMEP data - itself only covering the last 40 142 years, especially if up to 10 cm of the uppermost vegetation (possibly representing up to 20 143 vears of information) were not sampled. This would preclude both comparison with recent 144 EMEP data and any evaluation of the reliability of peat bogs for reconstructing deposition 145 chronologies.

146 | If Miszczak et al. (2019) began sampling their peat record at the accumulating peat (i.e., 147 ignoring the living vegetation), it is also unlikely that their sampling resolution would allow 148 **comparison with emissions data. In European ombrotrophic peatlands, long-term mean peat** 149 accumulation rates (i.e. excluding surface vegetation growth) have been estimated to range 150 | between c. 0.18 to 1 mm yr⁻¹ (e.g. Gorham, 1991; Mäkilä, 1997; Malmer and Wallén, 2004; 151 Pontevedra-Pombal, et al., 2017) depending on the vegetation and climate (e.g. Charman et 152 al., 2013; Pontevedra-Pombal et al., 2019). As a result, high-resolution sub-sampling and 153 dating are required to reconstruct decadal-scale atmospheric pollution records using peat 154 **cores and to properly compare them to 40 years of EMEP data (i.e. at best 4 cm of peat** 155 \vert accumulation, surface vegetation excluded if we take a 1mm $yr⁴$ peat accumulation rate). 156 The low sub-sampling resolution of Miszczak et al. (2019) and their removal of surface 157 vegetation, combined with limited radionuclide dating preclude any assessment of recent Pb 158 **deposition from their study.**

159

160 **3.2. Interpreting data in elemental ratios, enrichment factors (EFs) and metal**

161 **accumulation rates,**

162 Reconstructing Pb contamination in ombrotrophic peatlands requires an understanding of 163 how peat bogsthey respond to environmental change. This is because changes in Pb 164 concentrations may result from changes in the peat bog density/accumulation rate rather 165 | than changes in the extent of contamination. - Miszczak et al. (20192020) seem are perhaps 166 unaware of theseis important considerations. In the following section which follows we outline 167 the importance of understanding density changes in peat records and the need to 168 **takeconsider into account** the variability in natural Pb from aeolian mineral dust deposition.

169 We provide a brief overview of They therefore misinterpret data published in 170 Vleeschouwer et al. (2009a) and instead assume that those authors incorrectly interpreted 171 | their own data. Consequently, Miszczak et al. (2019) appear to misinterpret 172 **comprehend: 1) the effect of the density record presented in De Vleeschouwer et al. (2009b)** 173 and why the density and accumulation rates change within the Slowinskie Blota peat profile, 174 and 2) techniques to account for such changes, allowing and to reliably use Pb to be reliably 175 used as a tracer of past anthropogenic activity.

176

177 *3.2.1. Density, accumulation rate and Pb concentration*

178 Within The importance of understanding and accounting for changes in peat in density and 179 dust input are illustrated by De Vleeschouwer et al., (2009a,b) in their study of the Slowinskie 180 Blota ombrotrophic peatland in (Poland). That peatland is located close to those examined by 181 | Miszczak et al. (2020). In the the Slowinskie Blotathis peatland core, a section of higher than 182 average bulk density was present between 50 and 35 cm depth (De Vleeschouwer et al., 183 2009a,b). It corresponds to the timing of the Little Ice Age, when colder temperatures 184 promoted a decrease in peat accumulation rates coeval with increased windiness and dune 185 activity (i.e. increased aeolian lithogenic inputs; De Vleeschouwer et al., 2009b). This 186 combination of reduced organic accumulation rates and increased dust input resulted in $\frac{1}{n}$ 187 thean increase in bulk density in the bog. The result on theeffect of these changes $-Pb$ 188 **concentration profile** was an apparent increase in Pb concentration within the peat profile. 189 Although part of this Pb increase is attributable to increased pollution in the Industrial 190 Revolution, the majority of the Pb increase results from the decrease in peat accumulation 191 causing an apparent increase in pollution Pb contamination accumulation. This occurs 192 because that section of the peat profile represents a greater period of time $\frac{by}{y}$ 193 eomparisonthan the sections below 50 cm depth or above 35 cm depth. In addition, 194 increased dust inputs during the drier conditions of the Little Ice Age mean there was an 195 | increase in natural Pb input during that period. Miszczak et al. (2020) incorrectly incorrectly 196 assumed the increase in Pb in the Slowinskie Blota peatland at that time resulted from the 197 | movement of Pb from higher in the peat profile (i.e., Pb mobility in the bog). But, by 198 appropriately accounting for the change in density and increased dust input during the Little 199 Ice Age (i.e. using EFs accumulation rates and isotopic ratios, which are discussed in the 200 next section), De Vleeschouwer et al. $\frac{1}{12}$ (2009a, b) demonstrated the maximum Pb 201 concentration in the Slowinskie Blota record occurred at AD 1960-70s, and not between 50 202 and 35 cm depth as the raw Pb concentration data would suggest. Thus The maximum Pb 203 contamination therefore coincided precisely with maximum Pb emissions from leaded 204 gasoline, just prior to Pb being banned and phased out beginning in the 1980s (e.g. Pacyna 205 and Pacyna, 2000). De Vleeschouwer et al., $(2009a,b)$ therefore showed that the Pb 206 **contamination record in Slowinskie Blota**, matches the known history of anthropogenic Pb 207 emissions in Europe clearly demonstrating that it is not related to any post-depositional 208 mobility.₇ The approach to accurately reconstruct Pb contamination (as separate from total 209 Pb concentrations) is outlined in the following section.

 $\overline{ }$ -owing largely to the decrease in peat accumulation in addition to displaying an increase in **the concentrations of pollutant elements from the Industrial Revolution. Indeed, the maximum Pb concentration in the Slowinskie Blota record occurred at AD 1960-70s, coinciding with** 213 maximum Pb emissions from leaded gasoline, just prior to Pb being banned and phased out **beginning in the 1980s (e.g. Pacyna and Pacyna, 2000). It has therefore nothing to do with** \parallel any post-depositional Pb migration as claimed by Miszczak et al. (2019).

216

217 *3.2.2. Pb as a tracer of past anthropogenic activity*

218 Since the pioneering paper of Lee and Tallis (1973), practices have developed to ensure the 219 accurate use of trace metal data to reconstruct past environmental pollution. It has been 220 demonstrated that using concentration data to reconstruct past anthropogenic activity is 221 | problematic because, as shown above, peat accumulation $\frac{1}{2}$ and $\frac{1}{2}$ accumulation rates alter total 222 Pb concentrationsalter total Pb concentrations. A. Additionally, because the rate of dust 223 deposition in bogs (from wind erosion of soils) varies in time, it is necessary to separate 224 natural Pb in dust from anthropogenic Pb-within any studied bog sediments. $\overline{-}$ using EFs

- 252 $\overline{}$ SI1) demonstrating thatof Pb is being largely immobile in ombrotrophic peatlands and 253 showsing peatlands to be excellent recorders of anthropogenic activities.
- 254

255 **4. CONCLUSIONS**

256 | Due to their atmospheric-fidelity ombrotrophic peatlands have been extensively used to study 257 current and past patterns of atmospheric metal contamination and metal use, in particular for 258 Pb. Despite their utility there are some key considerations required when constructing 259 contamination histories from peat bogs. The aim of this comment was to highlight some 260 these considerations. The impetus for this arose from the recent paper by Miszczak et al 261 (2020) who used nonstandard sampling and analysis techniques and, as a result, came to 262 what we consider to be erroneous conclusions. Additional discussion on the effect of pH on 263 **Pb mobility and the relationship between peat age and the history of Pb contamination is** 264 **provided in the supplementary material.** 265 Cver the past 40 years a large number of many investigators have developed or applied a 266 range of analysis and sampling techniques necessary to construct metal contamination 267 <u>records in ombrotrophic peatlands. These approaches include undertaking high resolution</u> 268 sampling and dating, including sampling the living vegetation of the surface of peat-bogs 269 and the use of short--lived radionuclides (such as $^{210}_{6}Pb$) to accurately reconstruct metal 270 contamination over the past \sim 100 years or less. They also include calculating enrichment 271 factors (EFs), elemental ratios, or using accumulations rates (as opposed to raw metal 272 concentration data) to take account of changes in peat density/growth rates and changes in 273 natural metal input. Studies of Pb contamination have also benefitted from the use of Pb

274 isotopes to decipher emission sources at a regional to continental scale. Many of these steps

275 are also necessary when accurately determining contaminate loads and patterns in other 276 environments including in ice, lakes and soils and within direct atmospheric samples, 277 **aAlthough.** Wwe therefore wish to reiterate the particular value of ombrotrophic peats for

278 reconstructing atmospheric metal contamination into chronologies due to their wide distribution 279 and high atmospheric-fidelity. We maintain that 40 years of literature demonstrate that Pb is **Formatted:** Superscript

Formatted: Not Highlight

Formatted: English (United Kingdom)

ACKNOWLEDGEMENTS

We are thankful to P.G. Appleby, R. Bindler, J. Sonke, B. Smieja-Krol and three anonymous reviewers for their constructive comments.

REFERENCES

- Allan, M., Le Roux, G., De Vleeschouwer, F., Bindler, R., Blaauw, M., Fagel N., 2013. High-resolution reconstruction of atmospheric deposition of trace metals and metalloids since AD 1400 recorded by ombrotrophic peat cores in Hautes-Fagnes, Belgium. Environ. Pollut. 178, 381-394.
- Appleby, P.G., 2001. Chronostratigraphic techniques in recent sediments, in Last, W.M. and Smol, J.P. (Eds.), Tracking Environmental Change Using Lake Sediments: Basin Analysis, Coring, and Chronological Techniques, Vol. 1, Kluwer Academics, pp. 171-203.
- Appleby, P.G., Oldfield, F., 1978. The calculation of lead-210 dates assuming a constant rate of supply of unsupported ²¹⁰Pb to the sediments. Catena 5, 1-8.
- Appleby, P.G., Shotyk, W., Fankauser, A., 1997. Lead-210 age dating of three peat cores in the Jura Mountains, Switzerland. Water Air Soil Pollut. 100, 223-231.
- Bindler, R. 2011. Contaminated lead environments of man: reviewing the lead isotopic evidence in sediments, peat, and soils for the temporal and spatial patterns of atmospheric lead pollution in Sweden. Environ. Geochem. Health 33, 311-329

Bindler, R., Klarqvist, M., Klaminder, J., Förster, J. 2004. Does within-bog variability of mercury and lead constrain reconstruction of absolute deposition rates from single peat records? The example of Store Mosse, Sweden. Gobla Biogechem. Cycles 18, GB3020.doi:10.1029/2004GB002270.

- Broder, T, Biester, H , 2017. Linking major and trace element concentrations in a headwater stream to DOC release and hydrologic conditions in a bog and peaty riparian zone. Appl.ied Geochemistry Geochem. 87, 188-201,
- Charman, D., Beilman, D.W., Blaauw, M., Booth, R.K., Brewer, S., Chambers, F.M., Christen, J.A., Gallego-Sala, A., Harrison, S.P., Hughes, P.D.M., Jackson, S.T., Korhola, A., Mauquoy, D., Mitchell, F.J.G., Prentice, I.C., van der Linden, M., De Vleeschouwer, F., Yu, Z.C., Alm, J., Bauer, I.E., Corish, Y.M.C., Garneau, M., Hohl, V., Huang, Y., Karofeld, E., Le Roux, G., Loisel, J., Moschen, R., Nichols, J.E., Nieminen, T.M., MacDonald, G.M., Phadtare, N.R., Rausch, N., Sillasoo, Ü., Swindles, G.T., Tuittila, E.-S., Ukonmaanaho, L., Väliranta, M., van Bellen, S., van Geel, B., Vitt, D.H., Zhao, Y. 2013. Carbon-cycle implications of climate-driven changes in peat accumulation during the last millennium. Biogeosciences 10, 929-944.
- Cloy, J.M., Farmer, J.G., Graham, M.C., MacKenzie, A.B., Cook, G.T., 2008. Historical records of atmospheric Pb deposition in four Scottish ombrotrophic peat bogs: An isotopic comparison with other records from western Europe and Greenland. Glob. Biogeochem. Cycle 22, GB2016.
- Cloy, J.M., Farmer, J.G., Graham, M.C., MacKenzie, A.B. 2009. Retention of As and Sb in ombrotrophic peat bogs: records of As, Sb and Pb deposition at four Scottish sites. Environ. Sci. Technol. 43, 1756-1762.
- Davies, L.J., Appleby, P., Jensen, B.J.L., Magnan, G., Mullan-Boudreau, G., Noernberg, T., Shannon, B., Shotyk, W., van Bellen, S., Zaccone, C., Froese, D.G., 2018. High-resolution age modelling of peat bogs from
northern Alberta, Canada, using pre- and post-bomb ¹⁴C, ²¹⁰Pb and historical cryptotephra. Quat. Geochronol. 47,138-162.
- De Vleeschouwer, F., Fagel, N., Cheburkin, A., Pazdur, A., Sikorski, J., Mattielli, N., Renson, V., Fialkiewicz, B., Piotrowska, N., Le Roux, G., 2009a. Anthropogenic impacts in North Poland over the last 1300 years - A record of Pb, Zn, Cu, Ni and S in an ombrotrophic peat bog. Sci. Total Environ. 407, 5674-5684.
- De Vleeschouwer, F., Piotrowska, N., Sikorsky, J., Pawlyta, J., Cheburkin, A.K., Le Roux, G., Lamentowicz, M., Fagel, N., Mauquoy, D., 2009b. Multiproxy evidence of `Little Ice Age' palaeoenvironmental changes in a peat bog from northern Poland. Holocene 19, 625-637.
- De Vleeschouwer, F., Le Roux, G., Shotyk, W., 2010a. Peat as an archive of atmospheric metal pollution: the example of Pb in Europe. in: Jackson, S., Charman, D. (Eds.). Peatland. PAGES Newsletter, April 2010, vol.

Formatted: Indent: Left: 0 cm, First line: 0 cm

Formatted: English (United States)

18, 20-22.

- De Vleeschouwer, F., Chambers, F.M., Swindles, G.T., 2010b. Coring and sub-sampling of peatlands for palaeoenvironmental research. in: De Vleeschouwer, F., Hughes, P., Nichols, J.,Chambers, F.M. (Guest Eds.), A review of protocols in peat palaeoenvironmental studies. Mires and Peat, vol. 7, article 1, 1-10.
- Farmer, J.G., Eades, L.J., Atkins, H., Chamberlain, D.F., 2002. Historical trends in the lead isotopic composition of archival *Sphagnum* mosses from Scotland (1838-2000). Environ. Sci. Technol. 36, 152-157.
- Farmer, J.G., Graham, M.C., Yafa, C., Cloy, J.M., Freeman, A.J. and MacKenzie, A.B., 2006, Use of ²⁰⁶Pb/²⁰⁷Pb ratios to investigate the surface integrity of peat cores used to study the recent depositional history and geochemical behaviour of inorganic elements in peat bogs, Glob. Planet.. Change.*,* 53, 240-248.
- Fiałkiewicz-Kozieł, B., Łokas, E., Gałka, M., Kołaczek, P., De Vleeschouwer, F., Le Roux, G., Smieja-Król B. (2020). Influence of transboundary transport of trace elements on mountain peat geochemistry (Sudetes, Central Europe). Quat. Sci. Rev. 230, 106162.
- Givelet, N., Le Roux, G., Cheburkin, A., Chen, B., Frank, J., Goodsite, M., Kempter, H., Krachler, M., Noernberg, T., Rausch, N., Rheinberger, S., Roos-Barraclough, F., Sapkota, A., Scholz, C., Shotyk, W., 2004. Suggested protocol for collecting, handling and preparing peat cores and peat samples for physical, chemical, mineralogical and isotopic analyses. J. Environ. Monit. 6, 481-492.
- Goodsite, M.E., Rom, W., Heinemeier, J., Lange, T., Ooi, S., Appleby, P.G., Shotyk, W., Van der Knapp, W.O.,
Lohse, C., Hansen, T.S., 2001. High-resolution AMS ¹⁴C dating of post-bomb peat archives of atmospheric pollutants. Radiocarbon 43**,** 453-473.
- Gorham, E., 1991. Northern peatlands: role in the carbon cycle and probable responses to climatic warming. Ecol. Appl. 1, 182-195.
- Hansson, S.V., Kaste, J.M., Chen, K., Bindler, R., 2014. Beryllium-7 as a natural tracer for short-term downwash in peat. Biogeochemistry 119, 329-339.
- Hansson, S.V., Tolu, J., Bindler, R. (2015). Downwash of atmospherically deposited trace metals in peat and the influence of rainfall intensity: an experimental test. Sci. Total Environ. 506, 95-101.
- Hughes, P.D.M., Mauquoy, D., Barber, K.E. Langdon, P.G., 2000. Mire-development pathways and palaeoclimatic records from a full Holocene peat archive at Walton Moss, Cumbria, England. Holocene 10, 465-479.
- Kempter, H., Frenzel, B., 2007. The geochemistry of ombrotrophic *Sphagnum* species growing in different microhabitats of eight German and Belgian peat bogs and the regional atmospheric deposition**.** Water Air Soil Pollut. 184, 29.
- Kylander, M.E., Weiss, D.J., Peiteado Varela, E., Taboada Rodriguez, T., Martinez-Cortizas, A. 2006. Archiving anthropogenic lead pollution in ombrotrophic peatlands. in: Martini, P.I., Chestworth, W., Martinez-Cortizas, A. (Eds.), Peatlands: basin evolution and depository of records on global environmental and climatic changes. Elsevier, pp. 479-497.
- Lee, J., Tallis, J., 1973. Regional and historical aspects of lead pollution in Britain. Nature 245, 216-218.
- Le Roux, G., De Vleeschouwer, F., 2010. Preparation of peat samples for inorganic geochemistry used as palaeoenvironmental proxies. in: De Vleeschouwer, F., Hughes, P., Nichols, J., Chambers, F.M. (Guest Eds.), A review of protocols in peat palaeoenvironmental studies. Mires and Peat, vol. 7, article 4, 1-9.
- Li, C., Le Roux, G., Sonke, J., van Beek, P., Souhaut, M., Van der Putten, N., De Vleeschouwer, F., 2017. Recent
²¹⁰Pb, ¹³⁷Cs and ²⁴¹Am accumulation in an ombrotrophic peatland from Amsterdam Island (Southern Indian Ocean). J. Environ. Radioact. 175-176, 164-169.
- Longman, J., Veres, D., Finsinger, W., Ersek, V., 2018. Exceptionally high levels of lead pollution in the Balkans from the Early Bronze Age to the Industrial Revolution. Proc. Natl. Acad. Sci. U. S. A. 115, E5661-E5668.
- Mäkilä, M., 1997. Holocene lateral expansion, peat growth and carbon accumulation on Haukkasuo, a raised bog in southeastern Finland. Boreas 26,1-14.
- Malmer, N., Wallén, B., 2004. Input rates, decay losses and accumulation rates of carbon in bogs during the last millennium: internal processes and environmental changes. Holocene 14, 111-117.

bogs in northern Alberta, Canada reveal decades of declining atmospheric Pb contamination, Geophys. Res. Lett. 43, 9964-9974.

- Shotyk, W., Rausch, N., Nieminen, T.M., Ukonmaanaho, L., Krachler M., 2016b. Isotopic composition of Pb in peat and porewaters from three contrasting ombrotrophic bogs in Finland: Evidence of chemical diagenesis in response to acidification. Environ. Sci. Technol. 50, 9943-9951.
- Smieja-Król, B, Fiałkiewicz-Kozieł, B, Sikorski, J, Palowski, B., 2010. Heavy metal behaviour in peat a mineralogical perspective. Sci. Total Environ. 408, 5924-5931.
- Smieja-Król, B, Fiałkiewicz-Kozieł, B, Michalska, A, Krzykawski, T, Smołka-Danielowska, D., 2019. Deposition of mullite in peatlands of southern Poland: Implications for recording large-scale industrial processes. Environ. Pollut. 250, 717-727.
- Steinnes, E., Sjøbakk, T.E., 2005. Order-of-magnitude increase of Hg in Norwegian peat profiles since the outset of industrial activity in Europe. Environ. Pollut. 137, 365–370.
- Syrovetnik, K., Malmstrom, M.E., Neretnieks, I., 2007. Accumulation of heavy metals in the Oostriku peat bog, Estonia: Determination of binding processes by means of sequential leaching. Environ. Pollut. 147, 291-300.
- Twardowska, I., Kyziol, J., Goldrath, T., Avnimelech, Y., 1999. Adsorption of zinc onto peat from peatlands of Poland and Izrael. J. Geochem. Explor. 66, 387-405.
- Vile, M.A., Kelman Wieder, R., Novak, M., 1999. Mobility of Pb in Sphagnum-derived peat. Biogeochemistry 45, 35-52.
- Weiss, D., Shotyk, W., Kramers, J.D., Gloor, M., 1999. *Sphagnum* mosses as archives of recent and past atmospheric lead deposition in Switzerland. Atmos. Environ. 33, 3751-3763.
- Weiss, D., Rausch, N., Mason, T.F.D., Coles, B.J., Wilkinson, J.J., Ukonmaanaho, L., Arnold, T., Nieminen, T., 2007. Atmospheric deposition and isotope biogeochemistry of zinc in ombrotrophic peat. Geochim. Cosmochim. Acta 71, 3498-3517.

Comment on: "A novel approach to peatlands as archives of total cumulative spatial pollution loads from atmospheric deposition of airborne elements complementary to EMEP data: priority pollutants (Pb, Cd, Hg)"

by Ewa Miszczak, Sebastian Stefaniak, Adam Michczyński, Eiliv Steinnes and Irena Twardowska.

Authors

De Vleeschouwer F.^{1*}, Baron S.², Cloy J.M.³, Enrico M.⁴, Ettler V. ⁵, Fagel N.⁶, Kempter H.⁷, Kylander M.⁸, Li C.⁹, Longman J.¹⁰, Martinez-Cortizas A.¹¹, Marx S.¹², Mattielli N.¹³, Mighall T.¹⁴, Nieminen T.M.¹⁵, Piotrowska N.¹⁶, Pontevedra Pombal X.¹¹, Pratte S.¹⁷, Renson V.¹⁸, Shotyk W.¹⁹, Shuttleworth E.²⁰, Sikorski J.¹⁶, Stromsoe N.²¹, Talbot J.²², von Scheffer C.²³, D. Weiss 24,25 , Zaccone C. 25 , Le Roux G. 26

- 1. Instituto Franco-Argentino para el Estudio del Clima y sus Impactos (UMI IFAECI/CNRS-CONICET-UBA-IRD)
- 2. Laboratoire TRACES (CNRS, Université de Toulouse), France
- 3. Scotland's Rural College, Edinburgh, UK
- 4. Harvard John A. Paulson School of Engineering and Applied Science, Harvard University, Cambridge, MA, USA
- 5. Charles University, Prague, Czech Republic
- 6. AGEs, Département de Géologie, Université de Liège, Belgium
- 7. Welzheimer Str. 14, D-71566 Althuette, Germany
- 8. Department of Geological Sciences and the Bolin Centre for Climate Research, Stockholm University, Sweden
- 9. Geoscience Environnement Toulouse (CNRS-UPS-IRD-CNAP-CNES), France
- 10. School of Geography and the Environment, University of Oxford, UK
- 11. EcoPast (GI-1553), Universidad de Santiago de Compostela, Spain
- 12. GeoQuEST Research Centre, School of Earth, Atmospheric and Life Sciences, University of Wollongong, Australia
- 13. Laboratoire G-Time, Université Libre de Bruxelles, Belgium
- 14. School of Geosciences, University of Aberdeen, UK
- 15. Natural Resources Institute Finland Luke, Helsinki, Finland
- 16. Silesian University of Technology, Institute of Physics-CSE, GADAM Center, Gliwice, Poland
- 17. Department of Geography, School of Earth Sciences, Zhejiang University, Hangzhou, China
- 18. Research Reactor, University of Missouri, USA.
- 19. Department of Renewable Resources, University of Alberta, Edmonton, Canada
- 20. Department of Geography, University of Manchester, UK
- 21. College of Engineering, IT and Environment, Charles Darwin University, Australia
- 22. Département de Géographie, Université de Montréal, Canada
- 23. Institute for Ecosystem Research, Kiel University, Germany
- 24. Imperial College London, UK
- 25. 25 Princeton University, Princeton, USA
- 26. Department of Biotechnology, University of Verona, Italy
- 27. EcoLab, Université de Toulouse, CNRS, Toulouse, France

* Corresponding author: Dpto. de Ciencias de la Atmosfera y los Oceanos, FCEN, Universidad de Buenos Aires, Intendente Guiraldes 2160 - Ciudad Universitaria, Pabellon II - 2do. Piso, (C1428EGA) Ciudad Autonoma de Buenos Aires – Argentina. [fdevleeschouwer@cima.fcen.uba.ar](mailto: fdevleeschouwer@cima.fcen.uba.ar)

Abstract

A recent paper by Miszczak et al. (2020) examines metal contamination in mires in Poland and Norway. The authors conclude that lead (Pb) records in ombrotrophic peatlands cannot be used to reconstruct the chronological history of anthropogenic activities due to postdepositional mobility of the metal. We contest this general conclusion which stands in contrast with a significant body of literature demonstrating that Pb is largely immobile in the vast majority of ombrotrophic peatlands. Our aim is to reaffirm the crucial contribution that peat records have made to our knowledge of atmospheric Pb contamination. In addition, we re-iterate the necessity of following accepted protocols to produce reliable records of anthropogenic Pb contamination in environmental archives.

Keywords

Lead, immobility, ombrotrophic peatland, bog, geochemistry, enrichment factor, metal accumulation rate

1. INTRODUCTION

 Ombrotrophic peatlands are well-established archives of past atmospheric deposition of trace elements. After more than 40 years of investigation (see Table SI1), there is a consensus that Pb is largely immobile in ombrotrophic peatlands. We contend that some of 5 the conclusions reached by Miszczak et al (2020) are based on misinterpretation or incorrect sampling and data analysis approaches. To avoid such confusion, we seek to clarify here how ombrotrophic peatlands have allowed the reconstruction of past patterns of atmospheric metal contamination in the environment. We however do not aim to provide a complete review of Pb in mires, because the trophic status of a peatland is crucial to studies of past atmospheric metal deposition (see Supplementary Material for further details), we refer to a selection of studies relevant to ombrotrophic peatlands, also known as bogs. Due to their high atmospheric fidelity, ombrotrophic peatlands have special utility for reconstructing metal contamination records. This was also the type of peatland investigated by Miszczak et al. (2020).

2. **THE BEHAVIOUR OF LEAD IN OMBROTROPHIC PEATLANDS**

 While some studies have suggested that Pb can be mobilized in minerotrophic, riparian, drained or degraded peatlands (e.g. Syrovetnik et al., 2007; Smieja-Król et al., 2010, 2019; Rothwell, 2011; Broder and Biester, 2017) the majority of more than 40 years of literature suggests that Pb is largely immobile in pristine ombrotrophic peat profiles (e.g. De Vleeschouwer et al., 2010a; Marx et al., 2010, Shotyk et al., 2016a,b; Longman et al. 2018; Fiałkiewicz-Kozieł et al., 2020 and references therein). Although Miszczak et al. (2020) cite 23 literature to support their conclusion of Pb mobility in the bogs they examine, that literature pertains to minerotrophic or disturbed peatlands (e.g. Syrovetnik et al., 2007; Smieja-Król et al., 2010, 2019) and the processes that can promote Pb mobility in those systems are not applicable to ombrotrophic peatlands.

28 The similar elemental and isotopic trends encountered in ombrotrophic peat, lake sediment, ice and herbaria samples (e.g. Rosman et al., 1997; Weiss et al., 1999; Renberg et al., 2001; Farmer et al., 2002; Cloy et al. 2009; Bindler, 2011) and their agreement with anthropogenic emission patterns (e.g. Shotyk et al., 1998; Mighall et al., 2002; Kylander et al., 2006; De Vleeschouwer et al., 2009a; Marx et al., 2010; Bindler 2011; Cloy et al., 2008, 2009; Allan et al., 2013; Martínez-Cortizas et al., 2016) provide a body of evidence supporting the view that Pb is largely immobile in bogs. Significantly, stable Pb isotopes records from bogs have consistently been found to accurately reflect temporal variability in source signatures in numerous studies (e.g. in reference *op. cit.*). This would not be the case if post-depositional mobility/isotope mixing were taking place. Furthermore, in most ombrotrophic peat cores 210 Pb ages, as determined from the constant rate of supply (CRS) age-depth models (Appleby and Oldfield, 1978; Appleby, 2001), are in very good agreement with pollen chronological markers (Appleby et al., 1997), fallout radionuclide chronostratigraphic makers 41 (e.g. from ¹⁴C Bomb Pulse Curve, 137 Cs and 241 Am), and tephrochronology (e.g. Goodsite et al., 2001; Piotrowska et al. 2009; Li et al., 2017; Davies et al., 2018), providing *prima facie* evidence that Pb and its isotopes are largely immobile in bogs. Experimental studies lend further support to this (e.g. Vile et al., 1999; Novak et al., 2001). For example, Pb concentrations in the aqueous phase of ombrotrophic peatlands are low (<0.01% of total Pb), while the limited vertical water movement in bogs together with the size of the metal- containing particles in solution limits Pb redistribution (e.g. Shotyk et al., 2016b). Down- washing experiments have also demonstrated that Pb has limited mobility (Hansson et al. 2014, 2015). The limited mobility that may occur is not sufficient to compromise the use of Pb to reconstruct pollution histories over millennia. We note, however, the spatial distribution of Pb must be carefully addressed in cases where decomposition and compression integrate signals over longer (decadal and more) timespans (Bindler et al, 2004; Martinez Cortizas et al., 2012). Additionally, Pb behaviour in ombrotrophic peats has been demonstrated to differ from that of mobile elements such as Zn which, in contrast to Pb, displays evidence of vertical diffusion/advection as well as upward plant uptake (e.g. Shotyk 1988; Twardowska et al., 1999; Nieminen et al., 2002; Weiss et al., 2007).

 In summary, there is a significant body of evidence demonstrating that Pb is largely immobile in bog profiles (see Table S1 and Supplementary Material) that stands in contrast to the conclusions of Miszczak et al. (2020).

3. **HOW TO USE LEAD DATA TO ACCURATELY RECONSTRUCT HISTORICAL CONTAMINATION**

 In the following sections we outline what we consider to be the best practices to ensure accurate reconstruction of atmospheric Pb deposition. We also discuss appropriate approaches to use Pb pollution records constructed from peatlands to examine contaminant sources and to compare with emissions data. The approaches we outline are well established and have been described before (e.g. Givelet et al., 2004; De Vleeschouwer et al., 2010b). We hope that this overview corrects any misapprehensions arising from the approaches used by Miszczak et al. (2020).

3.1. Sampling and sub-sampling – Data resolution and geochronology

 Correct coring and sub-sampling protocols are important for accurately reconstructing metal contamination records from mires. The slow accumulation rate of ombrotrophic peatlands means that peat sections on the order of one vertical centimeter can represent decades of metal accumulation. For example, in European ombrotrophic peatlands, long-term mean peat accumulation rates (i.e. excluding surface vegetation growth) have been estimated to range 77 from c. 0.18 to 1 mm yr⁻¹ (e.g. Gorham, 1991; Mäkilä, 1997; Malmer and Wallén, 2004; Pontevedra-Pombal, et al., 2017) depending on the vegetation and climate (e.g. Charman et al., 2013; Pontevedra-Pombal et al., 2019). Because of this, it is commonplace for studies to both sample and date the living vegetation at the bog surface (e.g. Farmer et al., 2006; Kempter et al. 2007; Olid et al., 2008). This point was illustrated by Givelet et al. (2004) who stated: "*the historical record of atmospheric Pb … can depend to a large extent on the* *methods used to collect, handle, and prepare the samples for analysis*". As a result, high- resolution sub-sampling and dating are required to reconstruct decadal-scale atmospheric pollution records.

 Here Miszczak et al. (2020) compare metal contaminants in their peat records to European Monitoring and Evaluation Program (EMEP) data (annual trace metal emissions, [https://www.emep.int\)](https://www.emep.int/). Although potentially a very useful undertaking, their sampling approach unfortunately greatly reduces the utility of their comparison. This is because Miszczak et al. (2020) followed the coring protocol of Steinnes and Sjøbakk (2005) where "*Sphagnum moss … and other plant material growing on the surface were removed…before the coring, and the reference surface level is thus the interface moss/peat. The thickness of the Sphagnum layer, if present, was always less than 10 cm".* In other words, the authors removed the living/surface vegetation which is an integral part of the ombrotrophic peat deposit, potentially accumulating decades of information.

 Given the slow accumulation rates of bogs, the 40 years of EMEP data are likely, at best, to represent approximately 4 cm of peat accumulation (if surface vegetation is excluded and 98 assuming a 1mm $yr⁻¹$ peat accumulation rate). Therefore, the sampling resolution of Miszczak et al. (2020), where peats were subsampled in multi-centimeter increments, combined with their limited use of radionuclide dating, preclude any assessment of recent Pb deposition or comparison with EMEP data from their study. This is especially the case if surface vegetation were removed. Slow growth rates, combined with the demonstrated importance of surface vegetation in accumulating metal contaminants means that Miszczak et al. (2020) assumption that the peat/vegetation interface represents the year of coring (in that case 1999) is incorrect. Their approach therefore leads to large uncertainties in chronology and any inventory calculations performed thereafter. As previously stated, we consider comparing EMEP data with data from ombrotrophic peatlands to be a very worthwhile undertaking, but it requires high-resolution sub-sampling and dating, which is unfortunately not achieved by Miszczak et al. (2020).

3.2. Interpreting data in elemental ratios, enrichment factors (EFs) and metal

accumulation rates,

 Reconstructing Pb contamination in ombrotrophic peatlands requires an understanding of how they respond to environmental change. This is because changes in Pb concentrations may result from changes in the peat bog density/accumulation rate rather than changes in the extent of contamination. In the following section we outline the importance of understanding density changes in peat records and the need to consider the variability in natural Pb from aeolian mineral dust deposition. We provide a brief overview of techniques to account for such changes, allowing Pb to be reliably used as a tracer of past anthropogenic activity.

-
-

3.2.1. Density, accumulation rate and Pb concentration

 The importance of understanding and accounting for changes in peat density and dust input are illustrated by De Vleeschouwer et al. (2009a,b) in their study of the Slowinskie Blota ombrotrophic peatland (Poland). In this peatland, a section of higher than average bulk density was present between 50 and 35 cm depth (De Vleeschouwer et al., 2009a,b). It 127 corresponds to the timing of the Little Ice Age, when colder temperatures promoted a decrease in peat accumulation rates coeval with increased windiness and dune activity (i.e. increased aeolian lithogenic inputs). This combination of reduced organic accumulation rates 130 and increased dust input resulted in an increase in bulk density. The effect of these changes was an increase in Pb concentration within the peat profile. Although part of this Pb increase is attributable to increased pollution in the Industrial Revolution, the majority of the Pb increase results from the decrease in peat accumulation causing an apparent increase in 134 pollution Pb accumulation. This occurs because that section of the peat profile represents a greater period of time than sections below 50 cm depth or above 35 cm depth. In addition, increased dust inputs during the drier conditions of the Little Ice Age mean there was an increase in natural Pb input during that period. Miszczak et al. (2020) incorrectly assumed the increase in Pb in the Slowinskie Blota peatland at that time resulted from the movement

 of Pb from higher in the peat profile (i.e., Pb mobility in the bog). But, by appropriately 140 accounting for the change in density and increased dust input during the Little Ice Age (i.e. using EFs accumulation rates and isotopic ratios, which are discussed in the next section), De Vleeschouwer et al. (2009a) demonstrated the maximum Pb concentration in the Slowinskie Blota record occurred at AD 1960-70s, and not between 50 and 35 cm depth as 144 the raw Pb concentration data would suggest. The maximum Pb contamination therefore 145 coincided precisely with maximum Pb emissions from leaded gasoline, just prior to Pb being banned and phased out beginning in the 1980s (e.g. Pacyna and Pacyna, 2000). De Vleeschouwer et al. (2009a,b), match the known history of anthropogenic Pb emissions in Europe clearly demonstrating that it is not related to any post-depositional mobility. The approach to accurately reconstruct Pb contamination (as separate from total Pb concentrations) is outlined in the following section.

-
-

3.2.2. Pb as a tracer of past anthropogenic activity

 Since the pioneering paper of Lee and Tallis (1973), practices have developed to ensure the accurate use of trace metal data to reconstruct past environmental pollution. It has been demonstrated that using concentration data to reconstruct past anthropogenic activity is problematic because, as shown above, peat accumulation rates alter total Pb concentrations. Additionally, because the rate of dust deposition in bogs (from wind erosion of soils) varies, it is necessary to separate natural Pb in dust from anthropogenic Pb. Therefore, it is common practice to use metal to lithogenic element ratios, enrichment factors (EFs), or elemental mass accumulation rates (e.g. Shotyk et al., 1998; Le Roux et al., 2010; Allan et al., 2013) to reconstruct contamination histories. These approaches are important to avoid misinterpretations based on examining concentration data alone. Miszczak et al. (2020) did not apply these standard approaches. By comparison De Vleeschouwer et al. (2009a) use EFs and Pb accumulation rates (combined to Pb isotopes and high-resolution sampling and dating including the surface vegetation) to come to a different set of conclusions regarding the utility of Polish ombrotrophic peatlands for reconstructing Pb contamination histories.

 Consequently the latter represents a more accurate picture of the extent of Pb contamination in Poland over the past 1400 years, demonstrating *i)* the Pb accumulation rate in the topmost centimeter of the peat is of the same order of magnitude as 2009 European Pb deposition (www.msceast.org, [www.emep.int\)](http://www.emep.int/), *ii)* the main sources of anthropogenic Pb are from metallurgy, coal and gasoline and, *iii)* the peak in Pb contamination matches the history of European Pb emissions. That work therefore provides another example amongst many others, (Table SI1) of Pb being largely immobile in ombrotrophic peatlands and shows peatlands to be excellent recorders of anthropogenic activities.

4. CONCLUSIONS

 Due to their fidelity ombrotrophic peatlands have been extensively used to study current and past patterns of atmospheric metal contamination and metal use, in particular for Pb. Despite their utility there are some key considerations required when constructing contamination histories from bogs. The aim of this comment was to highlight some these considerations. The impetus for this arose from the recent paper by Miszczak et al (2020) who used nonstandard sampling and analysis techniques and, as a result, came to what we consider to 183 be erroneous conclusions. Additional discussion on the effect of pH on Pb mobility and the relationship between peat age and the history of Pb contamination is provided in the supplementary material. Over the past 40 years many investigators have developed or applied a range of analysis and sampling techniques necessary to construct metal contamination records in ombrotrophic peatlands. These approaches include undertaking high resolution sampling and dating, including sampling the living vegetation of the surface of 189 bogs and the use of short-lived radionuclides (such as ^{210}Pb) to accurately reconstruct metal 190 contamination over the past $~100$ years or less. They also include calculating enrichment factors (EFs), elemental ratios, or using accumulations rates (as opposed to raw metal concentration data) to take account of changes in peat density/growth rates and changes in natural metal input. Studies of Pb contamination have also benefitted from the use of Pb isotopes to decipher emission sources at a regional to continental scale. Many of these steps

 are also necessary when accurately determining contaminate loads and patterns in other environments including in ice, lakes and soils and within direct atmospheric samples. We therefore wish to reiterate the particular value of ombrotrophic peats for reconstructing atmospheric metal contaminant chronologies due to their wide distribution and high fidelity. We maintain that 40 years of literature demonstrate that Pb is largely immobile in ombrotrophic peatlands (i.e. bogs) and that peat cores extracted from this type of mire represent reliable archives for reconstructing past natural changes in Pb deposition from 202 natural processes and anthropogenic activity. The approach is supported by experimental work and similar reconstructions of metal contamination in other environmental archives (herbarium samples, lake sediments, ice cores). We conclude by noting that reconstructions of Pb contamination from bogs provide unequivocal evidence of the global scale of atmospheric Pb contamination and a reliable record of the timing of changes in atmospheric deposition extending from pre-history until the present day.

ACKNOWLEDGEMENTS

We are thankful to P.G. Appleby, R. Bindler, J. Sonke, B. Smieja-Krol and three anonymous reviewers for their constructive comments.

REFERENCES

- Allan, M., Le Roux, G., De Vleeschouwer, F., Bindler, R., Blaauw, M., Fagel N., 2013. High-resolution reconstruction of atmospheric deposition of trace metals and metalloids since AD 1400 recorded by ombrotrophic peat cores in Hautes-Fagnes, Belgium. Environ. Pollut. 178, 381-394.
- Appleby, P.G., 2001. Chronostratigraphic techniques in recent sediments, in Last, W.M. and Smol, J.P. (Eds.), Tracking Environmental Change Using Lake Sediments: Basin Analysis, Coring, and Chronological Techniques, Vol. 1, Kluwer Academics, pp. 171-203.
- Appleby, P.G., Oldfield, F., 1978. The calculation of lead-210 dates assuming a constant rate of supply of unsupported 210 Pb to the sediments. Catena 5, 1-8.
- Appleby, P.G., Shotyk, W., Fankauser, A., 1997. Lead-210 age dating of three peat cores in the Jura Mountains, Switzerland. Water Air Soil Pollut. 100, 223-231.
- Bindler, R. 2011. Contaminated lead environments of man: reviewing the lead isotopic evidence in sediments, peat, and soils for the temporal and spatial patterns of atmospheric lead pollution in Sweden. Environ. Geochem. Health 33, 311-329
- Bindler, R., Klarqvist, M., Klaminder, J., Förster, J. 2004. Does within-bog variability of mercury and lead constrain reconstruction of absolute deposition rates from single peat records? The example of Store Mosse, Sweden. Gobla Biogechem. Cycles 18, GB3020.doi:10.1029/2004GB002270.
- Broder, T, Biester, H , 2017. Linking major and trace element concentrations in a headwater stream to DOC release and hydrologic conditions in a bog and peaty riparian zone. Appl. Geochem. 87, 188-201,
- Charman, D., Beilman, D.W., Blaauw, M., Booth, R.K., Brewer, S., Chambers, F.M., Christen, J.A., Gallego-Sala, A., Harrison, S.P., Hughes, P.D.M., Jackson, S.T., Korhola, A., Mauquoy, D., Mitchell, F.J.G., Prentice, I.C., van der Linden, M., De Vleeschouwer, F., Yu, Z.C., Alm, J., Bauer, I.E., Corish, Y.M.C., Garneau, M., Hohl, V., Huang, Y., Karofeld, E., Le Roux, G., Loisel, J., Moschen, R., Nichols, J.E., Nieminen, T.M., MacDonald, G.M., Phadtare, N.R., Rausch, N., Sillasoo, Ü., Swindles, G.T., Tuittila, E.-S., Ukonmaanaho, L., Väliranta, M., van Bellen, S., van Geel, B., Vitt, D.H., Zhao, Y. 2013. Carbon-cycle implications of climate-driven changes in peat accumulation during the last millennium. Biogeosciences 10, 929-944.
- Cloy, J.M., Farmer, J.G., Graham, M.C., MacKenzie, A.B., Cook, G.T., 2008. Historical records of atmospheric Pb deposition in four Scottish ombrotrophic peat bogs: An isotopic comparison with other records from western Europe and Greenland. Glob. Biogeochem. Cycle 22, GB2016.
- Cloy, J.M., Farmer, J.G., Graham, M.C., MacKenzie, A.B. 2009. Retention of As and Sb in ombrotrophic peat bogs: records of As, Sb and Pb deposition at four Scottish sites. Environ. Sci. Technol. 43, 1756-1762.
- Davies, L.J., Appleby, P., Jensen, B.J.L., Magnan, G., Mullan-Boudreau, G., Noernberg, T., Shannon, B., Shotyk, W., van Bellen, S., Zaccone, C., Froese, D.G., 2018. High-resolution age modelling of peat bogs from northern Alberta, Canada, using pre- and post-bomb ¹⁴C, ²¹⁰Pb and historical cryptotephra. Quat. Geochronol. 47,138-162.
- De Vleeschouwer, F., Fagel, N., Cheburkin, A., Pazdur, A., Sikorski, J., Mattielli, N., Renson, V., Fialkiewicz, B., Piotrowska, N., Le Roux, G., 2009a. Anthropogenic impacts in North Poland over the last 1300 years - A record of Pb, Zn, Cu, Ni and S in an ombrotrophic peat bog. Sci. Total Environ. 407, 5674-5684.
- De Vleeschouwer, F., Piotrowska, N., Sikorsky, J., Pawlyta, J., Cheburkin, A.K., Le Roux, G., Lamentowicz, M., Fagel, N., Mauquoy, D., 2009b. Multiproxy evidence of `Little Ice Age' palaeoenvironmental changes in a peat bog from northern Poland. Holocene 19, 625-637.
- De Vleeschouwer, F., Le Roux, G., Shotyk, W., 2010a. Peat as an archive of atmospheric metal pollution: the example of Pb in Europe. in: Jackson, S., Charman, D. (Eds.). Peatland. PAGES Newsletter, April 2010, vol. 18, 20-22.
- De Vleeschouwer, F., Chambers, F.M., Swindles, G.T., 2010b. Coring and sub-sampling of peatlands for palaeoenvironmental research. in: De Vleeschouwer, F., Hughes, P., Nichols, J.,Chambers, F.M. (Guest Eds.), A review of protocols in peat palaeoenvironmental studies. Mires and Peat, vol. 7, article 1, 1-10.
- Farmer, J.G., Eades, L.J., Atkins, H., Chamberlain, D.F., 2002. Historical trends in the lead isotopic composition of archival *Sphagnum* mosses from Scotland (1838-2000). Environ. Sci. Technol. 36, 152-157.
- Farmer, J.G., Graham, M.C., Yafa, C., Cloy, J.M., Freeman, A.J. and MacKenzie, A.B., 2006, Use of ²⁰⁶Pb/²⁰⁷Pb ratios to investigate the surface integrity of peat cores used to study the recent depositional history and geochemical behaviour of inorganic elements in peat bogs, Glob. Planet. Change*,* 53, 240-248.
- Fiałkiewicz-Kozieł, B., Łokas, E., Gałka, M., Kołaczek, P., De Vleeschouwer, F., Le Roux, G., Smieja-Król B. (2020). Influence of transboundary transport of trace elements on mountain peat geochemistry (Sudetes, Central Europe). Quat. Sci. Rev. 230, 106162.
- Givelet, N., Le Roux, G., Cheburkin, A., Chen, B., Frank, J., Goodsite, M., Kempter, H., Krachler, M., Noernberg, T., Rausch, N., Rheinberger, S., Roos-Barraclough, F., Sapkota, A., Scholz, C., Shotyk, W., 2004. Suggested protocol for collecting, handling and preparing peat cores and peat samples for physical, chemical, mineralogical and isotopic analyses. J. Environ. Monit. 6, 481-492.
- Goodsite, M.E., Rom, W., Heinemeier, J., Lange, T., Ooi, S., Appleby, P.G., Shotyk, W., Van der Knapp, W.O., Lohse, C., Hansen, T.S., 2001. High-resolution AMS¹⁴C dating of post-bomb peat archives of atmospheric pollutants. Radiocarbon 43**,** 453-473.
- Gorham, E., 1991. Northern peatlands: role in the carbon cycle and probable responses to climatic warming. Ecol. Appl. 1, 182-195.
- Hansson, S.V., Kaste, J.M., Chen, K., Bindler, R., 2014. Beryllium-7 as a natural tracer for short-term downwash in peat. Biogeochemistry 119, 329-339.
- Hansson, S.V., Tolu, J., Bindler, R. (2015). Downwash of atmospherically deposited trace metals in peat and the influence of rainfall intensity: an experimental test. Sci. Total Environ. 506, 95-101.
- Hughes, P.D.M., Mauquoy, D., Barber, K.E. Langdon, P.G., 2000. Mire-development pathways and palaeoclimatic records from a full Holocene peat archive at Walton Moss, Cumbria, England. Holocene 10, 465-479.
- Kempter, H., Frenzel, B., 2007. The geochemistry of ombrotrophic *Sphagnum* species growing in different microhabitats of eight German and Belgian peat bogs and the regional atmospheric deposition**.** Water Air Soil Pollut. 184, 29.
- Kylander, M.E., Weiss, D.J., Peiteado Varela, E., Taboada Rodriguez, T., Martinez-Cortizas, A. 2006. Archiving anthropogenic lead pollution in ombrotrophic peatlands. in: Martini, P.I., Chestworth, W., Martinez-Cortizas, A. (Eds.), Peatlands: basin evolution and depository of records on global environmental and climatic changes. Elsevier, pp. 479-497.
- Lee, J., Tallis, J., 1973. Regional and historical aspects of lead pollution in Britain. Nature 245, 216-218.
- Le Roux, G., De Vleeschouwer, F., 2010. Preparation of peat samples for inorganic geochemistry used as palaeoenvironmental proxies. in: De Vleeschouwer, F., Hughes, P., Nichols, J., Chambers, F.M. (Guest Eds.), A review of protocols in peat palaeoenvironmental studies. Mires and Peat, vol. 7, article 4, 1-9.
- Li, C., Le Roux, G., Sonke, J., van Beek, P., Souhaut, M., Van der Putten, N., De Vleeschouwer, F., 2017. Recent 210 Pb, 137 Cs and 241 Am accumulation in an ombrotrophic peatland from Amsterdam Island (Southern Indian Ocean). J. Environ. Radioact. 175-176, 164-169.
- Longman, J., Veres, D., Finsinger, W., Ersek, V., 2018. Exceptionally high levels of lead pollution in the Balkans from the Early Bronze Age to the Industrial Revolution. Proc. Natl. Acad. Sci. U. S. A. 115, E5661-E5668.
- Mäkilä, M., 1997. Holocene lateral expansion, peat growth and carbon accumulation on Haukkasuo, a raised bog in southeastern Finland. Boreas 26,1-14.
- Malmer, N., Wallén, B., 2004. Input rates, decay losses and accumulation rates of carbon in bogs during the last millennium: internal processes and environmental changes. Holocene 14, 111-117.
- Martinez-Cortizas, A., López-Merino, L., Bindler, R., Mighall, T., Kylander, M.E., 2016. Early atmospheric metal pollution provides evidence for Chalcolithic/Bronze Age mining and metallurgy in southwestern Europe. Sci. Total Environ. 545, 398-406.
- Martínez Cortizas, A., Peitedo Varela, E., Bindler, R., Biester, H., Cheburkin, A. 2012. Reconstructing historical Pn and Hg pollution in NW Spain using multiple cores from Chao de Lamoso bog (Xistral Mountains). Geochi. Cosmochim. Acta 82, 68-78.
- Marx, S.K., Kamber, B.S., McGowan, H.A. and Zawadzki, A. 2010. Atmospheric pollutants in alpine peat bogs record a detailed chronology of industrial and agricultural development on the Australian continent. Environ. Pollut. 158, 1615-1628.
- Mighall, T.M., Abrahams, P.W., Grattan, J.P., Hayes, D., Timberlake, S., Forsyth, S., 2002. Geochemical evidence for atmospheric pollution derived from prehistoric copper mining at Copa Hill, Cwmywtwyth, mid-Wales. Sci. Total Environ. 292, 69-80.
- Miszczak, E., Stefaniak, S., Michczyński, A., Steinnes, E., Twardowska, I., 2020. A novel approach to peatlands as archives of total cumulative spatial pollution loads from atmospheric deposition of airborne elements complementary to EMEP data: priority pollutants (Pb, Cd, Hg). Sci. Total Environ. 705, 135776.
- Novak, M., Zemanova, L., Voldrichova, P., Stepanova, M., Adamova, M., Pacherova, P., Komarek, A., Krachler, A., Prechova, E., 2011. Experimental evidence for mobility/immobility of metals in peat. Environ. Sci. Technol. 45, 7180-7187.
- Nieminen, T.M., Ukonmaanaho, L., Shotyk, W., 2002. Enrichments of Cu, Ni, Zn, Pb and As in an ombrothrophic peat bog near a Cu-Ni smelter in Southwest Finland. Sci. Total Environ. 292, 81-89.
- Olid, C., Garcia-Orellana, J., Martinez-Cortizas, A., Masqué, P., Peiteado, E., Sanchez-Cabeza, J.-A., 2008. Role of surface vegetation in ²¹⁰Pb-dating of peat cores.
- Pacyna, J.M., Pacyna, E.G., 2000. Atmospheric emissions of anthropogenic lead in Europe: improvements, updates, historical data and projections. GKSS report no. 2000/31, Geesthacht, Germany.
- Pontevedra-Pombal, X., Castro, D., Carballeira, R., Souto, M., López-Sáez, J.A., Pérez-Díaz, S., Fraga, M.I., Valcárcel, M., García-Rodeja, E., 2017. Iberian acid peatlands: types, origin and general trends of development. Mires and Peat 19, 1-19.
- Pontevedra-Pombal, X., Castro, D., Souto, M., Fraga, I., Blake, W.H., Blaauw, M., López-Sáez, J.A., Pérez-Díaz, S., Valcárcel, M., García-Rodeja, E., 2019. 10,000 years of climate control over carbon accumulation in an Iberian bog (southwestern Europe). Geosci. Front. 10, 1521-1533.
- Piotrowska, N., De Vleeschouwer, F., Sikorski, J., Pawlyta, J., Fagel, N., Le Roux, N., Pazdur, A., 2009. Intercomparison of radiocarbon bomb pulse and ²¹⁰Pb age models. A study in a peat bog core from North Poland. Nucl. Instrum. Methods Phys. Res. Sect. B: Beam Interact. Mater. Atoms 268, 1163-1166.
- Renberg, I. Bindler, R., Brännvall, M.L., 2001. Using the historical atmospheric lead-deposition record as a chronological marker in sediment deposits in Europe. Holocene 11, 511-516.
- Rosman, K.J.R., Chisholm, W., Hong, S., Candelone, J. P., Boutron, C.F., 1997. Lead from Carthaginian and Roman Spanish mines isotopically identified in Greenland ice dated from 600 B.C. to 300 A.D. Environ. Sci. Technol. 31, 3413-3416.
- Rothwell, J.J., Taylor, K.G., Evans M.G., Allott T.E.H., 2011. Contrasting controls on arsenic and lead budgets for degraded peatland catchment in Northern England. Environ. Pollut. 159, 3129-3133.
- Shotyk, W., 1988. Review of the inorganic geochemistry of peats and peatland waters. Earth-Sci. Rev. 25, 95- 176.
- Shotyk, W., Weiss, D., Appleby, P. G., Cheburkin, A. K., Frei, R., Gloor, M., Kramers, J.D., Reese, S., van Der Knaap, W.O., 1998. History of atmospheric lead deposition since 12,370¹⁴C yr BP from a peat bog, Jura mountains, Switzerland. Science 281, 1635-1640.
- Shotyk, W., Appleby, P.G., Bicalho, B., Davies, L., Froese, D., Grant-Weaver, I., Krachler, M., Magnan, G., Mullan-Boudreau, G., Noernberg, T., Pelletier, R., Shannon, B., van Bellen, S., Zaccone, C., 2016a. Peat bogs in northern Alberta, Canada reveal decades of declining atmospheric Pb contamination, Geophys. Res. Lett. 43, 9964-9974.
- Shotyk, W., Rausch, N., Nieminen, T.M., Ukonmaanaho, L., Krachler M., 2016b. Isotopic composition of Pb in peat and porewaters from three contrasting ombrotrophic bogs in Finland: Evidence of chemical diagenesis in response to acidification. Environ. Sci. Technol. 50, 9943-9951.
- Smieja-Król, B, Fiałkiewicz-Kozieł, B, Sikorski, J, Palowski, B., 2010. Heavy metal behaviour in peat a mineralogical perspective. Sci. Total Environ. 408, 5924-5931.
- Smieja-Król, B, Fiałkiewicz-Kozieł, B, Michalska, A, Krzykawski, T, Smołka-Danielowska, D., 2019. Deposition of mullite in peatlands of southern Poland: Implications for recording large-scale industrial processes. Environ. Pollut. 250, 717-727.
- Steinnes, E., Sjøbakk, T.E., 2005. Order-of-magnitude increase of Hg in Norwegian peat profiles since the outset of industrial activity in Europe. Environ. Pollut. 137, 365–370.
- Syrovetnik, K., Malmstrom, M.E., Neretnieks, I., 2007. Accumulation of heavy metals in the Oostriku peat bog, Estonia: Determination of binding processes by means of sequential leaching. Environ. Pollut. 147, 291-300.
- Twardowska, I., Kyziol, J., Goldrath, T., Avnimelech, Y., 1999. Adsorption of zinc onto peat from peatlands of Poland and Izrael. J. Geochem. Explor. 66, 387-405.
- Vile, M.A., Kelman Wieder, R., Novak, M., 1999. Mobility of Pb in Sphagnum-derived peat. Biogeochemistry 45, 35-52.
- Weiss, D., Shotyk, W., Kramers, J.D., Gloor, M., 1999. *Sphagnum* mosses as archives of recent and past atmospheric lead deposition in Switzerland. Atmos. Environ. 33, 3751-3763.
- Weiss, D., Rausch, N., Mason, T.F.D., Coles, B.J., Wilkinson, J.J., Ukonmaanaho, L., Arnold, T., Nieminen, T., 2007. Atmospheric deposition and isotope biogeochemistry of zinc in ombrotrophic peat. Geochim. Cosmochim. Acta 71, 3498-3517.

UMI-3351 INSTITUT FRANCO-ARGENTIN D'ETUDES SUR LE CLIMAT ET SES IMPACT INSTITUTO FRANCO-ARGENTINO DEL ESTUDIO DEL CLIMA Y SUS IMPACTOS

Subject: Conflict of Interest Statement

Buenos Aires, March 13th, 2020

To Science of the Total Environment Co-Editors in chief Damià Barceló and Jay Gan.

Dear Dr Barcelo and Dr Gan,

On Behalf of all my co-authors, I hereby declare I have no conflict of interest with any of the authors of Miszczak et al. (2019).

On the behalf of: Baron S. (CNRS, Toulouse, France), Cloy J.M. (Scotland's Rural College, Edinburgh, UK), Enrico M. (Harvard, USA), Ettler V. (Charles University Prague, Czech Republic), Fagel N. (U. Liège, Belgium), Kempter H., Kylander M. (U. Stockholm, Sweden), Li C. (GET, Toulouse, France), Longman J. (U. Oxford, UK), Martinez-Cortizas A. (U. Santiago de Compostela, Spain), Marx S. (U. Wollongong, Australia), Mattielli N. (U. Brussels, Belgium), Mighall T. (U. Aberdeen, UK), Nieminen T.M. (Natural Resources Institute, Helsinki, Finland), Piotrowska N. (GADAM Center, Poland), Pontevedra Pombal X. (U. Santiago de Compostela, Spain), Pratte S. (Zhejiang University), Renson V. (U. Missouri, USA), Shotyk W. (U. Alberta, Canada), Shuttleworth E. (U. Manchester, UK), Sikorski J. (GADAM Center, Poland), Talbot J. (U. Montreal, Canada), von Scheffer C. (U. Kiel, Germany), Weiss D. (Imperial College London, UK), Zaccone C. (U. Verona, Italy), Le Roux G. (CNRS, Toulouse, France)

Sincerely Yours,

François De Vleeschouwer

CNRS Researcher, Adjunct Director Instituto Franco-Argentino para el Estudio del Clima y sus Impactos UMI IFAECI, CNRS-CONICET-UBA-IRD, Buenos Aires, Argentina Tel: +54 11 5285-8458, fdevleeschouwer@cima.fcen.uba.ar

Supplementary Material [Click here to download Data in Brief: Supplementary Information to De Vleeschouwer et al_R2.docx](http://ees.elsevier.com/stoten/download.aspx?id=3258005&guid=95ef63e4-4f12-4716-85b1-8939a87ea545&scheme=1) **Supplementary Table [Click here to download Data in Brief: TABLE SI1_R2.xlsx](http://ees.elsevier.com/stoten/download.aspx?id=3258006&guid=9ddd4c85-32f9-4fb0-9322-9a1a86a952cb&scheme=1)**