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El Niño drought and tropical forest conversion synergistically determine mosquito development rate

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1 El Niño drought and tropical forest conversion synergistically determine mosquito 2 development rate

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Abstract

Extreme warming events can profoundly alter the transmission dynamics of mosquito-borne diseases by affecting mosquito life-history traits (e.g. survival, growth and reproduction). At local scales, temperatures are determined largely by vegetation structure and can be dramatically altered by drivers of land-use change (e.g. forest conversion). Disturbance activities can also hinder the buffering capacity of natural habitats, making them more susceptible to seasonal climate variation and extreme weather events (e.g. droughts). In experiments spanning three years, we investigated the interactive effects of tropical forest conversion and climate on fine-scale temperature, and the consequences for mosquito larval development. This study was conducted in the northern Malaysian Bornean state of Sabah using local *Aedes albopictus* mosquitoes; important vectors of dengue, chikungunya and Zika viruses. We demonstrate that variation in microclimates due to forest conversion dramatically increases development rates in *Ae. albopictus* mosquitoes. However, this effect was mediated by an El Niño Southern Oscillation (ENSO) drought event. In normal years, mean temperatures did not differ between land-use types, however mosquitoes reared in oil palm plantations typically emerged 2-3 days faster than in logged forests. During an ENSO drought, mean temperatures did differ between land-use types, but surprisingly this did not result in different mosquito development rates. Driving this idiosyncratic response may be the differences in daily temperature fluctuations between the land-use types that either push mosquito larvae towards optimal development, or over the thermal optimum, thereby reducing fitness. This work highlights the importance of

22 considering the synergistic effects of land-use and seasonal climate variations for predicting the
23 thermal response of a key mosquito life-history trait driving disease transmission dynamics.

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1 1. Introduction

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3 Extreme warming events can influence the distribution and dynamics of vector-borne disease
4 transmission (Messina *et al.* 2015; Vincenti-Gonzalez *et al.* 2018). Inter-annual climate
5 variations, such as the El Niño southern oscillation (ENSO), have been linked to outbreaks of
6 dengue (Vincenti-Gonzalez *et al.* 2018) and malaria (Hashizume *et al.* 2009; Chaves *et al.* 2012)
7 across the globe. Dramatic changes to local climates driven by land conversion (e.g.
8 deforestation) are also associated with altered transmission dynamics of mosquito-borne diseases
9 (Conn *et al.* 2002; Fornace *et al.* 2016). However, at both scales the relationships between
10 extrinsic drivers and patterns of disease remains unclear (e.g. Johansson *et al.* 2009),
11 highlighting the need for better understanding of the interactions between the environment,
12 mosquitoes, and pathogen transmission.

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14 Composite metrics of mosquito-borne disease transmission are typically derived using mosquito
15 functional or life-history traits (henceforth used synonymously). For example, vectorial capacity,
16 which estimates the number of potentially infectious bites arising from all mosquitoes biting a
17 single human on a single day, incorporates parameters relating to mosquito survival, adult
18 feeding behaviour and vector competence. Small changes in these key parameters can
19 profoundly alter transmission dynamics (Garrett-Jones 1964). As mosquitoes are small-bodied
20 ectotherms, these life-history traits are all sensitive to ambient temperature (Lyimo *et al.* 1992;
21 Delatte *et al.* 2009). The thermal sensitivity of a life-history trait is often summarised as a
22 nonlinear asymmetric curve, comprising a thermal optimum (T_{opt}) at which the performance rate
23 is maximised, and critical minimum (CT_{min}) and maximum (CT_{max}) temperatures, at which
24 performance is zero (Huey & Kingsolver 1989; Angilletta Jr & Angilletta 2009). Typically, the
25 thermal sensitivity of mosquito life history traits is assessed experimentally, under constant
26 temperature conditions (Alto & Juliano 2001; Bayoh & Lindsay 2003). In contrast, wild

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3 27 mosquitoes encounter temperature fluctuations as great as 20°C in many normal transmission
4 28 settings (Paaijmans *et al.* 2009).
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8 30 Drivers of extreme warming events, such as ENSO and land-use change increase daily
9 31 temperature fluctuations even further (Luskin & Potts 2011). Importantly, regional climate and
10 32 local weather interact non-linearly to determine the temperatures to which mosquitoes are
11 33 exposed (Stenseth *et al.* 2003). At local scales, vegetation cover modifies solar radiation, air and
12 34 soil temperature, rainfall, air humidity and wind, to create a microclimate (Aussenac 2000). In
13 35 forests, dense canopies absorb relatively high amounts of solar radiation producing cool, less
14 36 variable microclimates (Hardwick *et al.* 2015; Kovács *et al.* 2017). Disturbance activities that
15 37 reduce canopy cover (e.g. selective logging or forest conversion) profoundly alter microclimates,
16 38 increasing mean and maximum temperatures and decreasing humidity (Luskin & Potts 2011;
17 39 Meijide *et al.* 2018). Importantly, the effects of vegetation cover on microclimate are mediated
18 40 largely by the general climate (Aussenac 2000) and different land-use types vary in their
19 41 buffering capacity against extreme climate events (e.g. ENSO).
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31 43 Identifying the effects of microclimate on mosquito population dynamics and pathogen
32 44 transmission is important. A growing body of work has found that fluctuating temperatures differ
33 45 considerably in their effects on mosquito life-history traits when compared to equivalent constant
34 46 mean temperatures (Lambrechts *et al.* 2011; Carrington *et al.* 2013a; Carrington *et al.* 2013b). In
35 47 general, fluctuations at lower mean temperatures act to speed up rate processes, whereas the
36 48 opposite occurs at high mean temperatures (Paaijmans *et al.* 2010). As a consequence, models
37 49 that use averaged temperatures and fluctuations collected over coarse spatiotemporal scales may
38 50 fail to accurately predict the effects on mosquito life-history, and consequently on the dynamics
39 51 of mosquito-borne disease transmission. This has been demonstrated for larval development
40 52 under laboratory conditions, where large fluctuations around a low mean temperature reduced
41 53 development time by approximately five days compared to constant temperature at the same
42 54 mean (Carrington *et al.* 2013a). Given that mosquito population dynamics are determined in part
43 55 by the rate at which new adult mosquitoes are produced (Garrett-Jones 1964), and that larval
44 56 development time is linked to a suite of adult traits relevant to vectorial capacity (Alto *et al.*
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2008; Araújo & Gil 2012; Zirbel *et al.* 2018), understanding how this life-history trait responds to ecologically realistic temperature fluctuations is critical to effective vector and disease control.

In a series of experiments conducted across a tropical agro-forest landscape, we investigated the effects of temperature variation on mosquito life-history. We used local Asian tiger mosquitoes, *Aedes albopictus*, a species native to Southeast Asia but that has spread rapidly throughout the globe, and is a vector for over 22 known arboviruses, including dengue, chikungunya and Zika viruses. Specifically, we ask: 1) how does the conversion of tropical forest to cropland alter temperatures to which mosquito larvae are exposed; 2) how does an ENSO drought affect these temperatures, and 3) how do land-use change and climate interact to affect mosquito larval development?

2. Methods

Sampling sites

Studies were carried out annually at the Kalabakan Forest Reserve (4°33'N, 117°16'E) in Malaysian Borneo, between January and April 2016-2018. The site forms part of the Stability of Altered Forest Ecosystems (SAFE) project, a large-scale deforestation and forest fragmentation experiment comprising dipterocarp tropical forest and oil palm plantations (Ewers *et al.* 2011). Forested sites are grouped into 14 sampling blocks, ranging from undisturbed primary forest to salvage logged forests, the latter of which are in the process of being converted into a fragmented agricultural landscape. The salvage logged forest sites have undergone multiple rounds of selective logging where medium to large commercial timber trees were removed once in the 1970s and one to three times from 1990-2008 (mean aboveground biomass (ABG) of trees with DBH ≥ 10 cm across all forest plots = 95.4 ± 72.8 t/ha (\pm standard deviation, SD; Pfeifer *et al.* 2016; Riutta *et al.* 2018). Differences in available timber and topography have resulted in a high degree of fine-scale spatial variation in logging intensity, and canopy cover ranges from 69.9 to 75% in salvage logged sites (Pfeifer *et al.* 2015). Average tree height in logged forests is 8.9 m \pm 2.31 (SD), and tree density is 336 ± 55 trees ha⁻¹ (Ewers 2018). Oil palm plantations were established as monocultures in 2006 (Ewers *et al.* 2011), and have considerably lower biomass

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3 19 than the forest sites (ABG = 38.1 ± 25.2 t/ha (\pm SD); Pfeifer *et al.* 2016). The plantations are
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5 20 characterised by sparse understory vegetation (123 ± 8.32 trees ha^{-1}) and low, open canopies
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7 21 (average tree height = 1.36 ± 1.26 m, SD; Ewers 2018). Forest and plantation sites are separated
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9 22 by approximately 15 km, but have been selected to minimise variation in altitude (mean = 460
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11 23 m; median = 460 m; interquartile range = 72 m). We randomly selected study sites from
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13 24 established SAFE project second-order sampling points in an oil palm plantation (OP2) and
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15 25 salvage logged forest (Fragment B). Sampling points are comprised of three points arranged 178
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17 26 m apart in a fractal pattern. For further details on the SAFE project sampling design, see Ewers
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19 27 *et al.* (2011). Climate in the region is typically aseasonal (Walsh & Newbery 1999) with
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21 28 occasional droughts that, are less frequent but more severe in Eastern Borneo than in other
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23 29 regions of the island, and are sometimes but not always associated with the positive phase of
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25 30 ENSO events (Walsh 1996). However, the 2015-2016 El Niño event was amongst the strongest
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27 31 in historical records (Null 2016; L'Heureux *et al.* 2017) and resulted in a severe drought in
28
29 32 Borneo (Meijide *et al.* 2018).

33 34 *Larval development experiments*

35 Mosquito eggs were collected every year from a single logged forest site ($4^{\circ}33'N$, $117^{\circ}16'E$) and
36
37 36 stored on paper towels until a sufficient number had been amassed. All eggs were less than two
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39 37 months old at the time of hatching. Eggs were hatched at the field camp by submerging egg
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41 38 papers in water for 24 hours, after which any unhatched eggs were discarded. The 1st instar
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43 39 larvae were distributed in rearing tanks (N = 3 tanks per site with 50 larvae per tank) to sites in
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45 40 oil palm plantation (N = 9) and in logged forest (N = 9) land-use types. Rearing tanks were
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47 41 comprised of 330 ml plastic cups (80 mm x 103 mm) covered with mesh cloth, and were placed
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49 42 in shaded areas at each sampling point. To minimise their risk of being knocked over by wildlife,
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51 43 tanks were buried ~3 cm into the ground. Rearing tanks were visited every 1-2 days and larvae
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53 44 were provisioned with approximately five *Tetra Cichlid Colour*TM fish pellets every two days.
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55 45 Pupae were removed and hatched at the field camp, and the number of development days (taken
56
57 46 as days from egg hatching to emergence) and the sex of emerging adults were recorded.
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59 47 Experiments were repeated once per year between 2016 and 2018.

53 54 55 48 56 57 49 *Temperature data*

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3 50 Air temperature was recorded every 30 minutes for the duration of each experiment, using a Data
4 51 Logger EL-USB-2 (LASCAR electronics, Salisbury). Loggers were secured to small wooden
5 52 stakes at a height of 5 cm, and placed with the rearing tanks at each sampling point. The loggers
6 53 were shaded from direct sunlight and rain with plastic plates suspended approximately 30 cm
7 54 above each post.
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14 1 3. Analysis

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16 3 *Land-use mediated temperature*

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18 5 Temperature data were filtered to remove outliers due to instrument failure. Outliers were
19 6 identified as points falling outside 1.5x the interquartile range of daily temperatures for each site
20 7 (See supplementary material (S1) for summary of outliers). All analyses were carried out in R
21 8 Version 3.5 (R Core Team 2014, <http://www.R-project.org>). Generalised linear models with
22 9 Gaussian errors (package ‘lme4’; link = log; Bates *et al.* 2012) were used to investigate the
23 10 effects of land-use type and ENSO drought event on mean temperature and daily temperature
24 11 range, and Gamma errors were used for minimum and maximum temperature. Both land-use
25 12 type and ENSO drought were treated as categorical variables. Models selection was carried out
26 13 using stepwise selection and Chi-squared statistic. Post-hoc significance tests using the glht
27 14 function (package = ‘multcomp’; Hothorn *et al.* 2008) were used to compare treatment effects
28 15 and their interactions.
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35 17 *Probability of mosquito emergence*

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37 19 Mean mosquito development times were compared between the different land-use types for an
38 20 ENSO and non-ENSO years using Wilcoxon-Mann-Whitney tests. To investigate the effects of
39 21 temperature and land-use type on mosquito larvae development, time-to-event models (Cox
40 22 proportional hazards; package ‘survival’; Terry 2012) were used. Proportional hazards models
41 23 are a semi-parametric regression method that analyse the effect of explanatory variables on a
42 24 hazard rate, defined as the instantaneous risk of an event occurring, given that it has not occurred
43 25 up until that time. Here, event was defined as day of adult mosquito emergence, taken as the
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3 26 number of days from hatching of eggs. For this study, higher hazard rates thus denote earlier
4 27 mosquito emergence times.
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8 29 Due to high correlation between ENSO and the temperature variables, we retained ENSO alone
9 30 in the model along with the uncorrelated land-use type, and mosquito sex. Model selection was
10 31 achieved using Akaike Information Criterion (AIC) and Chi-squared tests were used to ensure
11 32 that hazard functions were proportional over time for each treatment.
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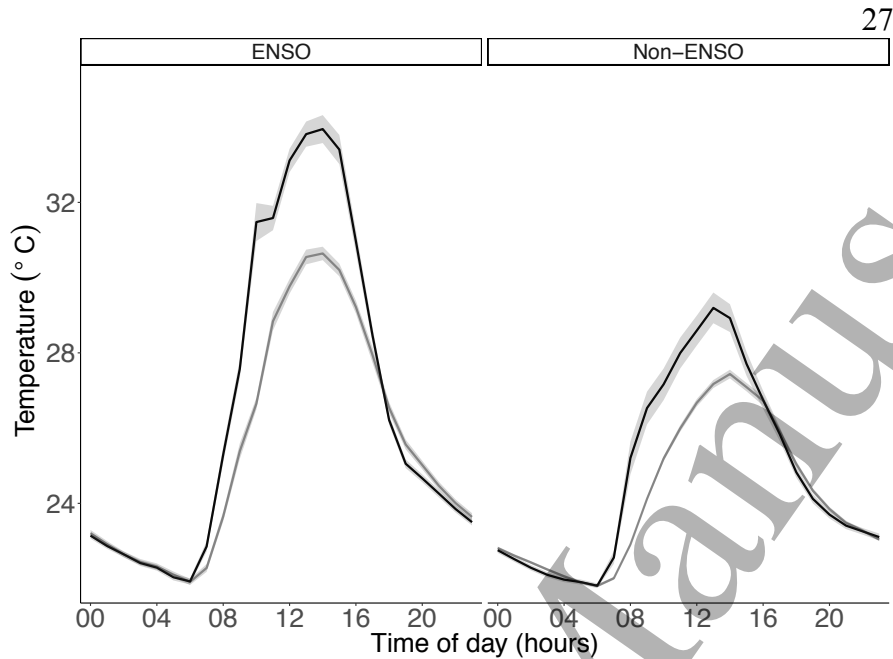
17 1 4. Results

18 2 19 3 20 4 21 5 *Effect of ENSO drought and land-use type on temperature* 22 6 23 7

24 5 Land-use type and ENSO drought both significantly affected the temperature to which mosquito
25 6 larvae were exposed (Fig 1). In non-ENSO periods, average daily temperatures in logged forest
26 7 and oil palm plantation sites did not differ significantly (GLM, $F = 12.37$, $p = 0.57$; Fig 2A).
27 8 While the average daily temperature did not differ during this period, the daily temperature
28 9 ranges at oil palm sites were greater (Fig 2A) due to higher temperature fluctuations (GLM, $F =$
29 10 13.17 , $p = 0.003$; Fig 1). Maximum temperatures were higher in oil palm plantation than in
30 11 logged forest sites during the ENSO period (GLM, $F = 6.41$, $p = 0.02$), however minimum
31 12 temperatures did not differ significantly between land-use types (GLM, $F = 1.11$, $p = 0.38$) in
32 13 either period.
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43 15 During the ENSO drought, average daily temperatures increased significantly across all land-use
44 16 types (GLM, $F = 59.36$, $p < 0.001$). However, temperatures in oil palm plantations were higher
45 17 and increased by more than twice as much than in logged forest, resulting in a higher average
46 18 daily temperature (Fig 2A). Again, daily temperature range was significantly greater in oil palm
47 19 plantations than in logged forests (GLM, $F = 13.17$, $p = 0.003$). Average maximum temperatures
48 20 were significantly greater in oil palm plantation than in logged forest, with oil palm plantations
49 21 experiencing peak temperatures of 7°C warmer on average than in logged forest (GLM, $F = 5.73$,
50 22 $p = 0.03$). Daily minimum temperatures did not differ by land-use type during the ENSO drought
51 23 (GLM, $F = 0.83$, $p = 0.38$). In both land-use types, average maximum temperatures were
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24 significantly greater during the ENSO drought than during non-drought periods (GLM, $F =$
 25 44.81, $p < 0.001$).



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 29 **Figure 1. Daily diurnal temperature cycles across the two land-use types.** Mean hourly temperature in logged
 30 forest (grey line) and oil palm plantations (black line) during a non-ENSO and ENSO period. Shaded areas denote
 31 standard errors.

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 37 *Effect of temperature on development time of Aedes albopictus mosquito larvae*

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 42 Mosquito larval development was significantly affected by local temperature and by ENSO (Fig
 43 2B; Table 1). During non-drought periods, the duration of larval development was longer in
 44 logged forests (Mann-Whitney U, $W = 119450$, $p < 0.001$). In logged forests development was
 45 16.8 ± 0.29 (SE) days for females and 14.2 ± 0.25 (SE) days for males, while in oil palm average
 46 development was 14.6 ± 0.25 (SE) days for females and 11.4 ± 0.22 (SE) days for males. The
 47 regional climatic shift associated with the ENSO drought decreased the average emergence time
 48 for mosquitoes reared in both land use types. During the ENSO drought there was no difference
 49 in the development time of larvae reared in different land use types (Mann-Whitney U, $W =$
 50 7282 , $p = 0.15$). In logged forests females took 10.9 ± 0.21 (SE) days to emerge and males took

51 9.29 ± 0.20 (SE) days. In oil palm plantations females took 10.3 ± 0.25 (SE) days and males took
52 9.09 ± 0.25 (SE) days to emerge.

53
54 Results of the Cox proportional hazards model supported the interacting effects of land-use type
55 and ENSO drought on mosquito emergence times (Table 1). Mosquito larvae developing during
56 an ENSO drought, that survive to adulthood, experienced a 7-fold increase in the hazard rate,
57 leading to earlier emergence times compared to those in a non-drought period (Cox Regression, z
58 = 16.42, $p < 0.001$). Developing in an oil palm plantation also significantly increased the hazard
59 rate of mosquito emergence, with mosquitoes 1.99 ± 0.08 times more likely to emerge per day
60 than in logged forest (HR = 1.99; Cox Regression, $z = 8.79$, $p < 0.001$). Emergence times were
61 also strongly influenced by sex, with male mosquitoes emerging earlier with a daily emergence
62 probability of 2.0 ± 0.07 times that of females (Cox Regression, $z = 10.42$, $p < 0.001$). There was
63 also a significant interaction between ENSO drought and land-use, such that the daily probability
64 of emergence of mosquitoes in oil palm plantations was 0.6 ± 0.15 times less during the drought
65 (Cox Regression, $z = -3.85$, $p = 0.0001$).

Factors	β	$e(\beta)$	$se(\beta)$	z	p
ENSO	1.96	7.08	0.12	16.42	<0.001**
Land-use type (Oil palm)	0.69	1.99	0.08	8.79	<0.001**
Sex (Male)	0.71	2.02	0.07	10.42	<0.001**
ENSO * Land-use type (Oil palm)	-0.59	0.55	0.15	-3.85	0.0001**

68 **Table 1. Coefficients of Cox proportional hazards survival analysis estimating effects of ENSO, sex and land-**
69 **use type on the timing of adult mosquito emergence.** Regression coefficients and hazard ratios (exponentiated
70 coefficients) are shown. Positive coefficients imply higher risk; negative coefficients imply the opposite. P values
71 indicate the significance of the coefficients using a Wald z statistic.

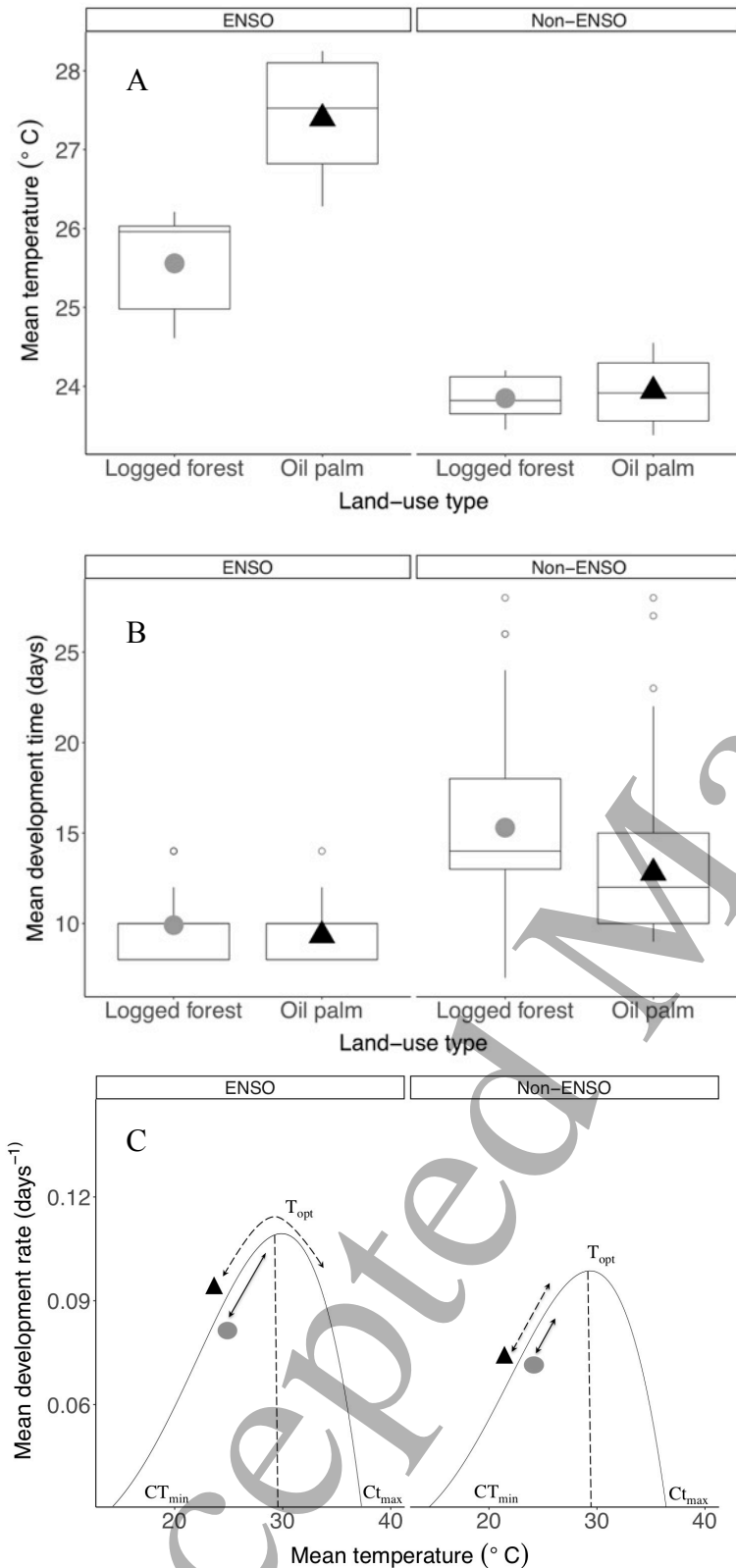


Figure 2. Summary of land-use effects on microclimate and on mosquito larval development. (A) Mean temperatures in logged forest and oil palm plantations; (B) Mean development time of *Ae. albopictus* larvae (measured as number of days from eggs hatching to adults emerging), vertical boxplot lines represent values within 1.5x the interquartile range of the upper and lower quartiles, horizontal boxplot lines indicate the quartile values, large points denote means for logged forest (grey circles) and oil palm (black diamonds), small points are outliers; (C) Conceptual thermal performance curve demonstrating how temperature fluctuations in logged forest (grey circles) and oil palm plantations (black diamonds) may affect larval development rate (CT_{min} = critical minimum, CT_{max} = critical maximum and T_{opt} = thermal optimum).

2 5. Discussion

4 *5.1 Logged forests buffer the effects of ENSO drought on local temperature conditions*

6 The results of this study indicate that oil palm plantations experience more severe changes in
7 local temperature than logged forests during an ENSO drought event. The greater buffering
8 capacity of logged forests has been identified in a number of other studies (Frey *et al.* 2016;
9 Meijide *et al.* 2018), and is related to greater canopy closure (Renaud *et al.* 2011) and higher leaf
10 area index (Hardwick *et al.* 2015). However, in contrast with a previous study, mean daily
11 temperature in the logged forest and oil palm plantation sites did not differ significantly in non-
12 drought years (Meijide *et al.* 2018). Where average midday temperatures are compared, the
13 differences in mean temperature between forest and oil palm can be even larger (Luskin & Potts
14 2011; Sabajo *et al.* 2017). The similarity in temperature observed in this study is likely due to the
15 maturity of the oil palm plantations sampled, as older plantations have more closed canopies and
16 more complex understory vegetation structure (Luskin & Potts 2011), and to the history of
17 selective logging which has reduced canopy cover in the forest sites (Ewers *et al.* 2011).

19 *5.2 Land-use and ENSO synergistically affect mosquito development*

21 The effects of land-use type on mosquito emergence times are surprising, given the difference in
22 their temperatures. In normal years, logged forests and oil palm plantations did not differ
23 significantly in mean temperature, however mosquitoes had considerably higher hazard rates in
24 oil palm plantations compared to logged forest. During an ENSO drought, when mean
25 temperatures and daily temperature ranges did differ significantly, mosquito hazard rates were
26 almost identical between the sites. It is worth reiterating that for this study, the hazard rate refers
27 only to adult emergence times and not to mosquito survival. It is possible that other microclimate
28 factors may influence mosquito larval development. For example, relative humidity is expected
29 to be greater in logged forest than in oil palm (Hardwick *et al.* 2015), which could decrease the
30 surface tension of water in the experimental pots (Pérez-Díaz *et al.* 2012) and thereby impact
31 mosquito pupation success (Murdock *et al.* 2017).

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3 33 An alternative explanation for the absence of a land-use effect on development during the ENSO
4 34 drought is that the elevated temperatures in the oil palm plantation were high enough to cause
5 35 stress and reduce the fitness of developing mosquitoes (Feder & Hofmann 1999). Thermal
6 36 performance curves, which describe the effects of changes in body temperature on physiological
7 37 sensitivity and fitness, are non-linear, such that mean performance of a life-history trait under
8 38 fluctuating conditions may differ to the performance of that life-history trait under mean
9 39 temperature (Fig 2C; Martin & Huey 2008). Fluctuations that raise temperatures towards a
10 40 mosquito's optimum temperature are expected to increase performance relative to constant
11 41 temperature around the same mean. However, fluctuations that push temperatures beyond the
12 42 thermal optimum may result in a decline in performance (Huey & Stevenson 1979). As the slope
13 43 of the performance curve is steeper above the thermal optimum, increases in body temperature
14 44 beyond this are associated with a relatively rapid reduction in fitness (Angilletta Jr & Angilletta
15 45 2009).

16 46
17 47 This would explain the homogeneity of mosquito emergence rates during the ENSO drought
18 48 despite temperatures differing significantly between the land-use types. Under drought
19 49 conditions, mosquitoes in both logged forest and oil palm sites would have experienced
20 50 temperatures closer to their thermal optimum, resulting in the observed faster development rates
21 51 compared to normal years. However, temperature fluctuations in oil palm were 5°C greater than
22 52 in logged forest, potentially exposing mosquito larvae to deleterious temperatures for long
23 53 enough to reduce growth rates (Fig 2C). The temperature fluctuations could also explain the
24 54 difference in larval development in non-ENSO years, when mean temperatures are similar in the
25 55 land-use types, but greater daily temperature fluctuations in oil palm plantations result in faster
26 56 larval development times. It is worth reiterating here that our study included only one ENSO
27 57 event, and that the response of mosquito larvae may vary with the strength and frequency of
28 58 different drought events. Additionally, whilst we controlled for larval habitat in this study, these
29 59 sites may vary considerably under natural conditions in the two land-use types; other factors may
30 60 then be more dominant drivers of larval development rates.

31 61
32 62 A considerable body of work has been dedicated to elucidating the effects of temperature on the
33 63 life-histories of mosquitoes (Parker 1952; Lyimo *et al.* 1992; Alto & Juliano 2001; Delatte *et al.*

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3 64 2009; Paaijmans *et al.* 2011; Couret *et al.* 2014), however key knowledge gaps remain in our
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5 65 understanding of the thermal environment experienced by these small-bodied ectotherms under
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7 66 natural conditions. This study identifies synergistic effects of land-use and climate on the
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9 67 development rates of mosquitoes, highlighting the need to characterise environmental conditions
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11 68 at finer scales than is typically done (e.g. using regional weather stations) to fully understand
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13 69 how environmental change may drive mosquito population dynamics. Importantly, because
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15 70 vectorial capacity is a composite measure of multiple life-history parameters, each with
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17 71 potentially different thermal sensitivities, the effect of temperature on disease transmission will
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19 72 be integrated across these parameters (Martin & Huey 2008). For example, mosquito larval
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21 73 development rate is negatively correlated with adult body size (Mohammed & Chadee 2011),
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23 74 which in turn can have downstream effects on adult survival and feeding behaviour (Nasci &
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25 75 Mitchell 1994; Farjana & Nobuko 2013). Thus, although mosquitoes developing during the
26
27 76 ENSO drought emerged earlier than those in a non-ENSO year, they are likely to be smaller,
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29 77 have reduced adult survival (Nasci 1986), and therefore fewer opportunities for host contact over
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31 78 a lifetime. However, due to their limited teneral reserves smaller mosquitoes may take more
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33 79 frequent bloodmeals (Takken *et al.* 1998), which would increase host contacts (Farjana &
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35 80 Nobuko 2013). Importantly, microclimate is one of a many linked predictors of land-use change.
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37 81 Understanding how key environmental drivers (e.g. temperature) mediate these trade-offs in
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39 82 transmission ecology will be critical to prediction of disease transmission and effective disease
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41 83 control.

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43 85 6. Acknowledgments

44 86

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3 95 7. Authors contributions
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