

1 Bio-energy with carbon capture and storage (BECCS):
2 Opportunities for performance improvement

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8 **Abstract**

9 This study evaluates the performance of a 500 MW pulverised fuel BECCS sys-
10 tem. A performance matrix is developed to assess the opportunities for BECCS
11 performance improvement in terms of: energy efficiency, carbon intensity, and
12 pollutant emissions. The effect of fuel properties was analysed for variable (i)
13 coal type (high/medium sulphur content), (ii) biomass type (wheat straw and
14 wood chips), (iii) moisture content, and (iv) biomass co-firing proportion %. It
15 was observed that the co-firing of biomass increased the quantity and quality
16 of waste heat available for recovery from the exhaust gas. The opportunities to
17 improve energy efficiency in the BECCS system include enhancing heat recovery
18 and using high performance solvents for CO₂ capture, such as biphasic mate-
19 rials. Implementing these approaches increased the power generation efficiency
20 from 31%_{HHV} (conventional MEA system) to 38%_{HHV} (using “new solvent” with
21 heat recovery). Furthermore, power generation efficiency was found to influence
22 the carbon intensity on an annual basis and annual capacity (load factor) of
23 the BECCS system. Significant reductions to SO_x emissions were achieved by
24 increasing biomass co-firing % or using low sulphur coal.

25 *Keywords:* Biomass, bio-energy, BECCS, carbon capture and storage (CCS),
26 efficiency, co-firing, co-combustion, greenhouse gas removal (GGR)

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27 **1. Introduction**

28 *1.1. Bio-energy with carbon capture and storage (BECCS)*

29 Carbon capture and sequestration (CCS) and “negative emissions” technolo-
30 gies will play an essential role in achieving deep reductions in atmospheric CO₂
31 concentration [1]. There is growing interest in bio-energy with carbon capture
32 and storage (BECCS) as a promising negative emissions technology, and as a
33 means to meet global warming targets of below 2°C target [1] and 1.5°C set
34 by COP21 [2]. This highlights the importance of having BECCS as a CO₂
35 mitigation option.

36 The BECCS technology was first introduced for hydrogen production [3],
37 before the concept was adapted for “negative emissions” electricity generation
38 [4]. Over the lifetime of biomass growth, there is a net transfer of atmospheric
39 CO₂ into the biomass. The CO₂ arising from the combustion of this biomass is
40 captured and stored in geological formations, enabling the permanent removal
41 of CO₂ from the atmosphere [5, 6], and potentially achieving an overall negative
42 carbon balance [5, 6, 7, 8, 9, 10, 11, 12]. Biomass is generally considered a CO₂
43 neutral substitute for fossil fuels, where co-combustion has also been shown
44 to reduce the emissions of pollutants SO_x, NO_x and particulates [13, 14, 15].
45 Another important advantage of bio-energy with CCS is that it provides reliable
46 firm low carbon electricity, unlike intermittent renewable energy sources (IRES)
47 such wind or photovoltaic [16, 17]. The economic loss caused by power outages is
48 two orders of magnitude greater than the cost of electricity [17]. Therefore, this
49 emphasises the value of having firm capacity technologies to balance the use of
50 IRES in an electricity system. Bio-energy with CCS is recognised as a practical
51 and immediate approach to mitigating the use of coal and decarbonising the
52 electricity sector.

53 *1.2. Approaches to improve efficiency*

54 *1.2.1. Enhancement of fuel properties*

55 The fuel composition and properties of biomass differ significantly from coal
56 [18]. Biomass typically has lower HHV (higher heating value) and higher mois-

57 ture content in comparison with coal. Furthermore, co-firing biomass tends
58 to increase fouling and slagging in the boiler furnace [19, 20, 21, 22]. Conse-
59 quently, biomass co-firing tends to reduce the energy efficiency of the power plant
60 [22, 23, 24]. The CO₂ capture process imposes an additional energy penalty due
61 to heat requirements for solvent regeneration [25, 26]. Enhancements to energy
62 efficiency are required to minimise the marginal cost of electricity generation,
63 enabling operation of the power plant at higher load factor [27, 28]. Therefore,
64 BECC performance improvements would improve commercial viability of the
65 technology and encourage large scale deployment.

66 Coal blending to meet power plant requirements is conventional practice
67 [29, 30, 31]. Until recently, the main motivation for coal blending has been to
68 reduce cost and to utilise more readily available coal resources, *i.e.*, indigenous
69 coal. However, as the fuel markets change, the availability of indigenous coals is
70 declining, increasing the use of imported coals [31]. Fuel quality has an impact
71 on almost every aspect of power plant operation and performance. Boilers are
72 typically designed based on a fuel specification, where there is normally an
73 allowable range for important fuel properties [29] (*e.g.*, HHV, moisture content,
74 ash content and composition, grindability). Table 1 shows typical fuel property
75 requirements for a pulverised coal-fired power plant. Within these property
76 limits, the power plant is expected to produce full load [30]. However, deviation
77 from this design fuel specification can be detrimental to plant performance as
78 efficiency would reduce outside the recommended property limits. The use of
79 alternative fuels in power plants requires detailed evaluation of the resulting
80 impacts on performance and cost [30]. The following lists the impact of fuel
81 properties on certain plant performance [29, 30, 31]:

- 82 • **Fuel handling and storage:** heating value, moisture content, volatile
83 matter content, ash content and composition, grindability;
- 84 • **Pulverising/milling:** heating value, moisture content, volatile matter
85 content, ash content and composition;
- 86 • **Combustion performance in the boiler furnace:** moisture, ash,

87 volatile matter, heating value, particle size distribution;

88 • **Ash management/handling and influence on heat transfer:** heat-
89 ing value, moisture, ash content and composition;

90 • **Pollutant emissions:** sulphur content, nitrogen content.

91 Depending on how performance is impacted, modifications to equipment and
92 operations may be required. In the case of biomass co-firing, various power
93 plant configurations and modifications have been developed to improve biomass
94 co-firing performance [32, 33, 34, 35, 36]. To minimise the risk to normal boiler
95 operation during biomass co-firing, necessary plant modifications may include
96 separate fuel feeding systems, the addition of biomass-dedicated burners and
97 ash handling systems. For appropriate design of the combustion facility, it is
98 essential that the fuel characteristics (*e.g.*, moisture, ash, heating value, ignition
99 temperature) of the biomass are well understood [37].

100 Fuel blending to meet power plant specifications and requirements enables
101 the use of alternative fuels, whilst maintaining energy output and preventing
102 damage to the boiler. Additionally, fuel blending can be used as a means of com-
103 plying with emission regulations, notably sulphur or mercury emissions [31]. For
104 instance, SO_x emissions during the combustion of high sulphur bituminous coals
105 have been shown to reduce significantly by co-firing with fuels of low sulphur
106 content, *e.g.*, biomass [14, 38, 39, 40], or low sulphur coal [41]. Although cer-
107 tain biomass fuel properties reduce plant efficiency (*e.g.*, high moisture) [22, 24],
108 some fuel properties can be used to improve the overall combustion performance.
109 In comparison to pure coal, biomass typically has higher oxygen content and
110 greater volatile matter [23, 35], which improves the reactivity and ignition char-
111 acteristics of the fuel blend [31, 42], and reduces unburnt carbon [43].

112 Another approach to enhance performance is the development of biomass
113 pretreatment techniques which improve biomass fuel properties, thus reduc-
114 ing the negative impacts of fuel properties that cause slagging/fouling and re-
115 duced combustion performance. It is desirable to alter the physical properties
116 of biomass to match those of coal, minimising the need to modify fuel handling

Table 1: Typical coal quality requirements for a pulverised coal-fired power plant [30]. ar = as received, mf = moisture free, daf = dry and ash free. The ‘Typical limits’ are ranges commonly reported from both literature and survey of power plant operators, where those in the brackets indicate the outer limits that are acceptable under certain circumstances.

Parameter	Desired values	Typical limits
Heating value (MJ/kg ar)	High	Min 24–25 (23)
Moisture content (% ar)	4–8	Max 12
Volatile matter (% mf)	20–35 15–20	Min 20 (side-fired furnaces) Max 20 (down-fired furnaces)
Ash content (% mf)	Low	Max 15–20 (max 30)
Sulphur content (% daf)	Low	Dependent on local pollution regulations and flue gas desulphurisation capacity
Nitrogen content (% daf)	Low	(0.8–1.1), dependent on local pollution regulations and NO _x control measures
Chlorine content (% daf)	Low	Max 0.1–0.3 (max 0.5)
Hardgrove Grindability Index	High	Min 50–55 (min 39)

117 and combustion equipment [42]. Biomass pretreatment can involve mechanical,
118 thermal, chemical or biological processes.

119 The “densification” process increases the bulk density of biomass to improve
120 the efficiency of downstream steps, including storage, handling, supply/feed [44].
121 Biomass densification can be achieved through pelletising mills [45], briquette
122 presses [46], screw extruders or agglomeration (*i.e.*, binding powder particles)
123 [47]. The options for pretreatment include:

- 124 • **Leaching or washing:** removes undesirable chemical components that
125 cause ash deposition issues (*i.e.*, the silicates, chlorides and sulphates of
126 potassium and calcium), or corrosion (*i.e.*, acidic compounds formed from
127 chlorine and sulphur) [42].
- 128 • **Dry torrefaction:** generates biomass fuel with similar properties to coal,
129 *e.g.*, reduces moisture content, increases heating value, reduces propensity
130 to reabsorb H₂O, improves grindability/milling characteristics, particles
131 are more spherical, also size is smaller and more uniform [42, 44, 48]. Dur-
132 ing torrefaction, raw biomass is heated up to temperatures between 200
133 and 300°C[49]. The energy requirement is a function of the inlet biomass
134 moisture content at the inlet, torrefaction temperature and reaction time.
135 For an inlet biomass moisture content of 15%, the energy requirement has
136 been found to be around 90 kWh per ton of biomass [50]. However, this
137 value typically increases significantly with higher moisture content as the
138 drying heat requirement constitutes a large fraction of the total heat duty
139 for torrefaction.
- 140 • **Hydrothermal carbonisation or wet torrefaction:** end-product is
141 similar to dry torrefaction but wet torrefaction produces a solid with
142 greater energy density [51].
- 143 • **Steam explosion:** pressurisation with saturated steam causes physical
144 and chemical changes, the end-product has high density and low moisture
145 reabsorption [44]. During the process, heat is required to heat the biomass

146 and generate the steam at the selected temperature. The heat duty varies
147 with the temperature of the process and biomass type. For instance, the
148 heat duty for steam explosion of switchgrass is 150 kWh per ton of biomass
149 at 140°C, whereas wheat straw requires 210 kWh per ton of biomass at
150 180°C [52].

151 Each densification and pretreatment process vary in terms of energy consump-
152 tion, impacts on chemical and physical properties, as well as end-product quality
153 [47]. From an energy efficiency perspective, there will be a compromise between
154 the energy consumption of the pretreatment process and level of fuel enhance-
155 ment. To determine the value of biomass pretreatment, the impact of enhancing
156 specific fuel properties on the overall power plant performance needs to be as-
157 sessed carefully.

158 Blending or co-firing biomass together with coal can provide a fuel that meets
159 power plant requirements. The biomass reduces the emissions of pollutants
160 SO_X and NO_X, whereas coal increases the heating value to improve combustion
161 performance. Further improvements to fuel properties can be achieved through
162 biomass pre-treatment (*e.g.*, drying, densification). Detailed evaluation of the
163 effect of blended fuel properties and fuel enhancement is necessary to understand
164 biomass co-firing combustion behaviour, as well as the subsequent effect on
165 power plant performance.

166 Due to limitations on fuel specification (as seen in Table 1), thus, the
167 biomass-coal co-firing ratio is restricted in existing coal pulverized boilers. In
168 a recent report on biomass co-firing, the IEA Bioenergy recommended a maxi-
169 mum biomass co-firing ratio of 10% of the energy input, beyond which, negative
170 impacts on combustion performance and ash management may occur. However,
171 the co-firing ration can increase up to 40% if the biomass has been milled to
172 the right particle size [53]. Using the specifications of a relatively low moisture
173 biomass, this would translate in a maximum co-firing of around 60% in terms of
174 mass basis. This value can be considered as the maximum co-firing ratio on a
175 mass basis, beyond which, adjustments (*e.g.* move the flame detection system)

176 or design modifications (*e.g.* bring the ignition plane back to its initial position,
177 ash deposition protection measures) must be performed on the boiler technology
178 [53].

179 1.2.2. Process design for efficiency improvements

180 The CO₂ capture process requires low grade thermal energy for solvent re-
181 generation, typically sourced as saturated steam at ~ 3 bar [25, 54]. The steam
182 is predominantly supplied through steam extraction from the steam cycle of the
183 power plant, which imposes an efficiency penalty on the system [55, 56]. Several
184 options for steam extraction and steam cycle design have been developed to
185 minimise the efficiency penalty incurred due to CO₂ capture, these include:

- 186 • Steam extraction at the cross-over pipe between the intermediate pressure
187 (IP) and low pressure (LP) turbines [57, 58, 59, 60, 61];
- 188 • Extraction of steam from within the LP steam turbine at a suitable point
189 [62];
- 190 • Optimised designs for steam cycle retrofits [54, 63].

191 An alternative approach is to modify the process configuration of the CO₂
192 capture process to reduce the energy penalty and improve efficiency. Some stud-
193 ies focus on cost-effective stripper configurations to improve energy efficiency,
194 *e.g.* Karimi *et al.* (2011) [64], Van Wagener & Rochelle (2011) [65]. Various
195 configurations for CO₂ removal from natural gas-based power plant have also
196 been evaluated [66, 67, 68]. Amrollahi *et al.* (2012) combined lean vapour com-
197 pression with absorber inter-cooling to decrease energy consumption from 3.74
198 to 2.71 MJ/kg CO₂ [68], whereas Øi *et al.* (2014) reduced heat consumption
199 from 3.26 MJ/kg CO₂ (baseline) to 2.90 MJ/kg CO₂ by only using the vapour
200 recompression configuration [66]. In contrast, Zhang *et al.* (2016) focussed on
201 evaluating configurations that minimised capital expense. There are several
202 extensive reviews on configuration modifications for CO₂ capture in coal-fired
203 power plants [69, 70, 71, 72]. Of these process modifications, the split-flow

204 configurations shows promise, providing a 8.3% reduction to reboiler duty (full
205 load), thus, reducing the efficiency penalty substantially [73].

206 Process modifications can be implemented to use waste heat recovery and
207 further improve the overall energy efficiency [74]. There are three potential
208 approaches to recover waste energy:

- 209 1. Heat recovery or heat integration within the power plant to improve effi-
210 ciency;
- 211 2. Waste heat recovery from the CO₂ capture process to use in the power
212 plant and improve efficiency;
- 213 3. Heat recovery from the power plant to use in the CO₂ capture process,
214 reducing the energy penalty.

215 Examples of the first approach include flue gas heat recovery, which can be
216 utilised for fuel drying [75], or applied in a low pressure economiser to heat the
217 steam cycle condensate [76, 77, 78, 79]. Pfaff *et al.* (2010) [74] investigated
218 approaches to recover energy from the CO₂ capture process to enhance power
219 plant efficiency. Heat recovered from the stripper overhead condenser and the
220 CO₂ compressor intercoolers in the capture process were used to pre-heat the
221 power plant steam cycle condensate and combustion air [74]. Alternatively, flue
222 gas heat recovery can supply energy for solvent regeneration in the CO₂ capture
223 process [80, 81, 82, 83, 84]. For this approach, the amount of heat that is recover-
224 able from flue gas depends on: (i) the point along the pollution control pathway
225 (*e.g.*, exit of boiler, electrostatic precipitator, flue gas desulphurisation); and
226 (ii) fuel type and quality [55]. In a coal-fired power plant, the measured flue
227 gas temperature at the economiser outlet is $\sim 345^{\circ}\text{C}$ [85], whereas the solvent
228 regeneration temperature is typically 120°C [86, 87, 88]. All of these studies
229 investigated the implementation of efficiency enhancements for applications in
230 fossil fuel power plants. The moisture content of biomass can change signifi-
231 cantly, affecting the flue gas heat transfer rate [89] and flue gas volume flow
232 rate [30], which in turn influences power plant thermal efficiency [82]. There-
233 fore, an extensive performance assessment of biomass-specific power systems is

234 necessary.

235 1.3. Study objectives

236 To improve the energy efficiency of BECCS in this study, heat recovery from
237 boiler exhaust flue gas is used to supply heat for the CO₂ capture process. This
238 paper evaluates the performance of a 500 MW dedicated BECCS system under
239 different operating conditions (*e.g.*, variable fuel properties, different capture
240 solvents, and effect of heat recovery). The effects on combustion performance
241 from different: (i) coal types (high and medium sulphur content), (ii) biomass
242 types (wheat straw and clean wood chips with variable moisture content), and
243 (iii) biomass co-firing proportions percentage was demonstrated. Blending coal
244 with biomass provides the benefit of reducing pollutant emissions. Also, by vary-
245 ing fuel moisture content, the impact of drying biomass on plant performance
246 was investigated. Additionally, the effect of solvent technology (*e.g.*, different
247 heat duty) and biomass co-firing proportion on the system energy efficiency
248 and carbon intensity was evaluated. These results lead to the development of a
249 performance matrix which summarises the effect of key process parameters.

250 The remainder of the paper is structured as follows: (i) First the methodol-
251 ogy and modelling procedure is presented. (ii) The performance of a 500 MW
252 BECCS plant is evaluated, assessing the influence of fuel quality and biomass co-
253 firing proportion on the adiabatic flame temperature, SO_x and NO_x emissions,
254 heat recovery, plant efficiency and carbon intensity. (iii) A performance matrix
255 is developed to assess the opportunities for BECCS performance improvement
256 in terms of minimising pollutant emissions, enhancing power efficiency, reducing
257 carbon intensity. Lastly, (iv) the paper concludes with a discussion of future
258 directions for this research field.

259 2. Methodology

260 The modelling procedure used to assess performance of the 500 MW pul-
261 verised fuel BECCS system has been summarised below. Detailed equations
262 and descriptions of the models is available in Bui *et al.* (2017)[83].

263 *2.1. Power plant and CO₂ capture model*

264 The fuel types selected for this analysis included (i) high sulphur coal, (ii)
265 low sulphur coal, (iii) clean virgin wood chip, and (iv) wheat straw. The mois-
266 ture content of biomass fuels varied between 5% to 50%. The properties and
267 composition of these fuels is provided in Table 2.

268 Three CO₂ capture solvents were considered:

- 269 1. High heat duty scenario: MEA with a regeneration energy requirement of
270 3600 MJ/t_{CO₂} [90], and temperature of 120°C;
- 271 2. Medium heat duty scenario: industrial solvent Cansolv, which requires
272 2300 MJ/t_{CO₂} and 120°C in the reboiler [86, 87];
- 273 3. Low heat duty scenario: uses “new solvent” that operates at 2000 MJ/t_{CO₂}
274 and 80°C [83].

275 The “new solvent” heat duty is based on performance data for biphasic sol-
276 vent systems, which can achieve overall heat duty reduction of 30% compared
277 to conventional MEA systems [93]. The heat duty of 2900 MJ/t_{CO₂} has been
278 reported to be the attainable limits for MEA systems [88]. Therefore, a 30%
279 reduction to an energy requirement of 2000 MJ/t_{CO₂} is in accordance with the
280 state-of-the-art systems in literature [94, 95].

281 The ultra-supercritical 500 MW coal-fired power plant with 90% post-combustion
282 CO₂ capture was modelled in the Integrated Environment Controlled Model
283 (IECM) software [90]. The fuel and solvent data was implemented in the IECM
284 model to calculate the fuel firing flow rate and the net power output, which was
285 subsequently used to determine the power plant efficiency (%_{HHV}) and carbon
286 intensity (kg CO₂ emitted per MWh generated).

287 *2.2. Thermochemical combustion model*

288 The thermochemistry of the co-combustion of coal with biomass was mod-
289 elled in the software FactSage 7.0 [96, 97]. Table 3 summarises the coal and
290 biomass fuel blending scenarios modelled in FactSage, where the biomass co-
291 firing proportion was increased from 0% to 50% at increments of 5%. The fuel

Table 2: Fuel properties and composition for the coal and biomass types used for this analysis.

	High sulphur coal	Medium sulphur coal	Clean wood chips	Wheat straw
Reference	[90]	[91]	[91]	[14, 92]
HHV (MJ·kg ⁻¹ dry)	27.14	27.06	19.16	19.22
Fuel composition	wt %	wt %	dry wt %	dry wt %
C	63.75	64.6	50	48.7
H	4.5	4.38	5.4	5.7
O	6.88	7.02	42.2	39.1
Cl	0.29	0.023	0.02	0.32
S	2.51	0.86	0.05	0.01
N	1.25	1.41	0.3	0.6
Moisture	11.12	9.5	–	–
Ash	9.7	12.2	2.0	5.5
Ash composition	% ash	% ash	% ash	% ash
SiO ₂	46.8	50	43.1	56.2
Al ₂ O ₃	18.0	30.0	8.9	1.2
Fe ₂ O ₃	20.0	9.8	3.9	1.2
CaO	7.0	4.0	28.0	6.5
MgO	1.0	0.5	4.2	3.0
Na ₂ O	0.6	0.1	2.0	1.3
K ₂ O	1.9	0.1	5.5	23.7
TiO ₂	1.0	2.0	0.4	0.06
P ₂ O ₅	0.2	1.8	2.2	4.4
SO ₃	3.5	1.7	1.8	1.1
MnO	0	0	0	1.34

292 firing flow rates from IECM in tonnes per hour was used as the mass basis for
 293 the calculations in FactSage. To ensure complete combustion, the excess air
 294 coefficient (λ) of 1.3 was used, which maintain an O_2 concentration of $\sim 5 - 6\%$
 295 in the flue gas (in accordance with industrial practice).

Table 3: Coal and biomass co-combustion scenarios modelled in FactSage.

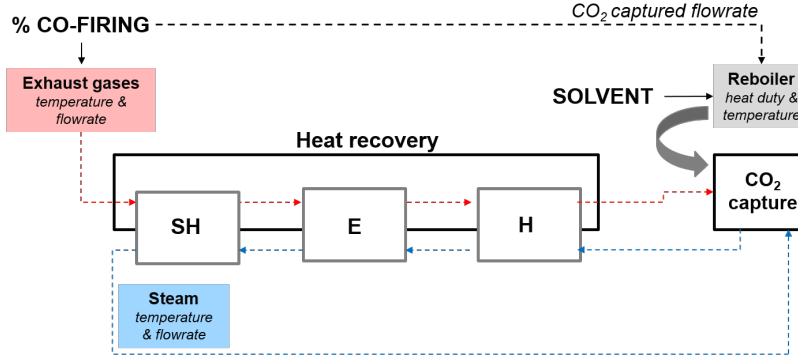
Scenario	Coal type	Biomass type
A	Medium sulphur coal	Wheat straw 5% moisture
B	Medium sulphur coal	Wheat straw 16% moisture
C	Medium sulphur coal	Wood chip 5% moisture
D	Medium sulphur coal	Wood chip 50% moisture
E	High sulphur coal	Wheat straw 5% moisture
F	High sulphur coal	Wheat straw 16% moisture
G	High sulphur coal	Wood chip 5% moisture
H	High sulphur coal	Wood chip 50% moisture

296 The combustion of each fuel blend was simulated from 200°C to the adiabatic
 297 flame temperature (the AFT was calculated in FactSage). The subsequent flue
 298 gas mixture was cooled from the AFT to 370°C (predicted flue gas temperature
 299 at the boiler exit in IECM). The objective was to calculate the following proper-
 300 ties for each biomass co-firing scenario: (i) AFT, (ii) SO_x and NO_x emissions,
 301 and (iii) exhaust gas properties at 370°C . These calculated exhaust gas prop-
 302 erties, *i.e.*, AFT, gas flow rate and specific heat capacity, were subsequently
 303 implemented in the heat recovery model.

304 2.3. Heat recovery model

305 A heat recovery model was developed in MATLAB; it was based on a heat
 306 exchanger system consisting of a heater, an evaporator and a super-heater [98].
 307 Using the input data from the FactSage analysis, the heat exchanger model
 308 (shown in Figure 1) was used to calculate the “recoverable heat”, which is the
 309 percentage of solvent heat duty recoverable from the boiler exhaust gases.

Figure 1: Schematic of the heat recovery model consisting of the super-heater (SH), evaporator (E) and heater (H).



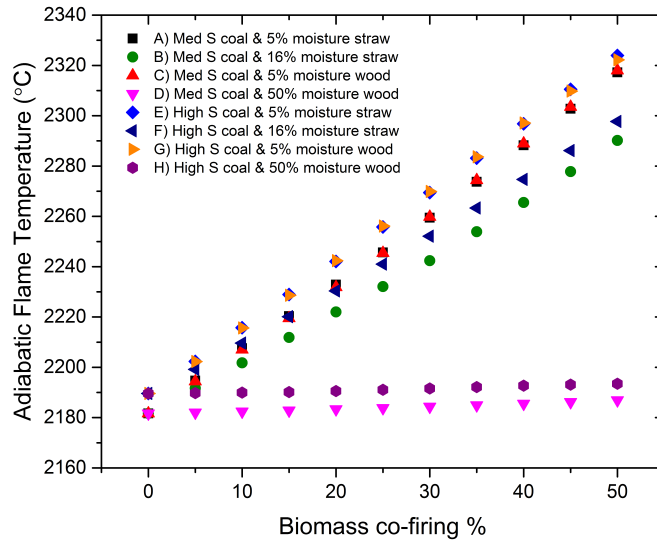
3. Performance evaluation of a 500 MW BECCS system

3.1. Adiabatic flame temperature (AFT)

Figure 2 illustrates that AFT generally increased with higher biomass co-firing proportions. For all fuel blends, the overall higher heating value (HHV) of the blended fuel reduced as biomass co-firing % increased, thus, the fuel firing rate increased in order to meet the specified power plant capacity (500 MW). The biomass moisture content had a significant influence on the degree at which AFT increased. The coal co-fired with biomass of low 5% moisture achieved the highest AFT increase. Compared to firing medium sulphur coal alone, 50% co-firing with 5% moisture wood/straw increased the AFT by 136°C (Scenario A and C). The straw with moderate moisture content of 16% also provided a substantial increase of 108°C in AFT at 50% biomass co-firing. In contrast, there was a marginal 4 – 5°C increase in AFT when co-firing coal with biomass of 50% moisture. The medium sulphur coal had higher ash content compared to high sulphur coal (shown in Table 2), consequently, co-firing the same biomass with medium sulphur coal generated lower AFT compared to high sulphur coal.

The fuel moisture and ash content had a strong impact on the combustion performance (*i.e.*, AFT), which is concordant with previous research [99]. The selection of fuels that reduce the moisture and ash content of the overall blend

Figure 2: Adiabatic flame temperature (AFT) for different co-firing scenarios with coal and biomass.



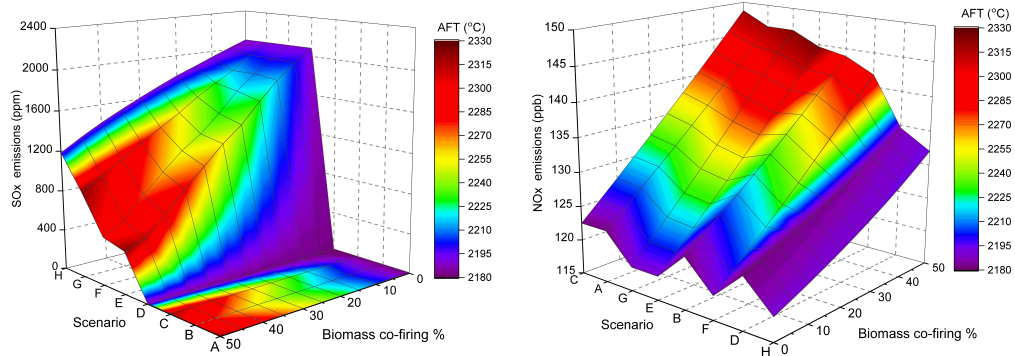
329 can provide higher AFT and potentially achieve greater flue gas temperatures.

330 3.2. Pollutant emissions

331 The SO_X and NO_X emissions have been represented as concentration of the
 332 exhaust flue gas at 370°C (from the FactSage analysis) in terms of parts per
 333 million (ppm) and parts per billion (ppb), respectively. Typically, the reduction
 334 in SO_X emissions during biomass co-firing are due to: (i) reduction in fuel
 335 sulphur content, (ii) presence of chemical compounds in ash that can absorb
 336 SO_2 (*e.g.*, alkali oxides) [14].

337 Figure 3 (left) shows that SO_X emissions reduced as the biomass co-firing
 338 proportion increased. Furthermore, the SO_X emissions from high sulphur coal
 339 co-firing were significantly greater than the medium sulphur coal scenarios. The
 340 scenarios involving high sulphur coal, Figure 4 (left), demonstrate SO_X emissions
 341 decreased linearly with higher biomass co-firing %, which can be attributed to
 342 the biomass (both the wood chip and wheat straw) having significantly lower

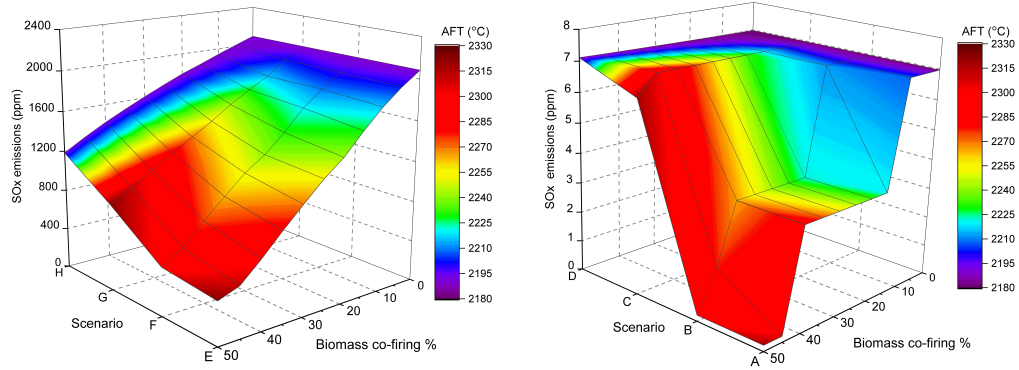
Figure 3: Emissions of SO_x (left) and NO_x (right) during the co-combustion of coal with biomass at different co-firing percentages. Refer to Table 3 for fuel blends in each Scenario.



343 sulphur content compared to coal. In contrast, biomass co-firing with medium
 344 sulphur coal resulted in a non-linear decrease of SO_x emissions, shown by Figure
 345 4 (right). As discussed by Bui *et al.* (2017) [83, 84], this non-linear (step-change)
 346 behaviour may be the result of equilibrium reaction shifts that occur at low fuel
 347 sulphur content. The FactSage analysis revealed that the alkali metal oxides in
 348 the ash also contributed to the reduction in SO_x emissions, specifically CaO ,
 349 MgO [14, 100], Na_2O and K_2O [101, 102].

350 Some experimental studies suggest increased biomass co-firing % reduced
 351 NO_x emissions [103, 104]. However, other studies have indicated high propor-
 352 tions of biomass co-firing can lead to NO_x emissions increasing [40, 105] or
 353 remaining unchanged [106] compared combustion of coal alone. The variation
 354 in the effect of biomass co-firing on NO_x emissions across different studies is
 355 due to the combustion conditions having a significant effect on the level of NO_x
 356 emissions during biomass co-firing [14, 15, 40, 107]. In particular, temperature
 357 and the presence of N_2 can significantly influence NO_x emission levels [108]. At
 358 combustion temperatures $>1300^\circ\text{C}$, thermal and prompt NO_x reaction path-
 359 ways can occur leading to NO_x formation from the N_2 in air [105, 109, 110].
 360 Figure 3 (right) shows AFT was above 1300°C , resulting in higher NO_x emis-
 361 sions due to increased formation from N_2 in air. In practice, combustion condi-

Figure 4: The SO_x emissions when biomass is co-fired with coal of high sulphur content (left) and medium sulphur content (right). Refer to Table 3 for fuel blends in each Scenario

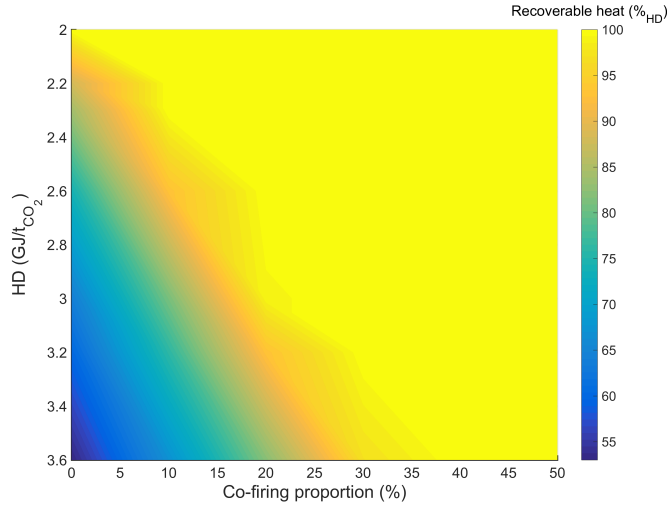


362 tions are optimised and controlled to ensure NO_x emissions are minimal (*e.g.*,
 363 temperature maintained below 1300°C , air staging, fuel staging) [105]. Future
 364 modelling work will need to consider such combustion control mechanisms to
 365 ensure the prediction of NO_x emissions are an accurate representation of an
 366 actual power plant combustion system.

367 3.3. Heat recovery

368 The heat recovery evaluation was conducted for co-combustion scenario of
 369 high sulphur coal with 16% moisture wheat straw (Scenario F). Figure 5 illus-
 370 trates the influence of solvent heat duty and biomass co-firing proportion on
 371 the recoverable heat (percentage of heat duty compensated by flue gas heat
 372 recovery). Of the three solvent scenarios at 0% biomass co-firing (*i.e.*, firing
 373 coal alone), heat recovery can only fulfil 100% of the regeneration energy re-
 374 quirements for the “new solvent” scenario (*i.e.*, heat duty of $2000 \text{ MJ}/t_{\text{CO}_2}$). To
 375 meet the specified power plant capacity, the fuel firing rate had to increase as
 376 the biomass co-firing % increased, as a consequence of biomass having a lower
 377 HHV compared to coal. As a result, the flow rate of the exhaust gas leaving
 378 the boiler increased with higher biomass co-firing proportions, thereby increas-
 379 ing the amount of recoverable heat. Once biomass co-firing proportion reaches

Figure 5: The fraction of solvent heat requirement that can be supplied from flue gas heat recovery as a function of the solvent heat duty (HD) and biomass co-firing percentage.

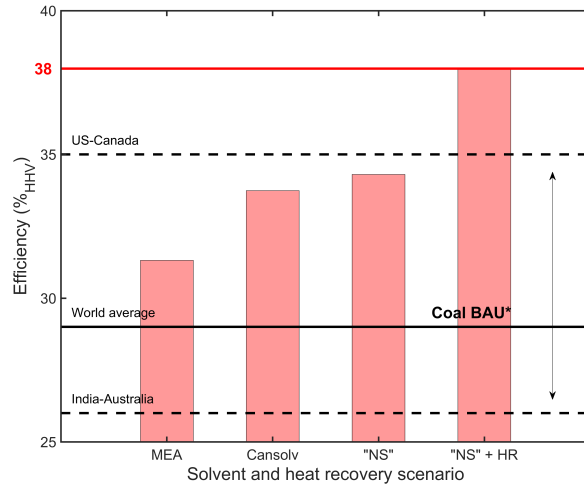


380 40%, heat recovery is capable of supplying all of the energy requirements in all
 381 three solvent scenarios. Therefore, flue gas heat recovery could enable operation
 382 of a BECCS plant without an efficiency penalty from CO₂ capture. However,
 383 this would be at the expense of some capital investment.

384 3.4. Plant efficiency

385 The plant efficiency for different scenarios at 50% biomass co-firing is illus-
 386 trated by Figure 6. The MEA system (baseline) had an efficiency of 31%_{HHV}.
 387 A significant increase in BECCS plant efficiency was achieved by using a more
 388 advanced CO₂ capture solvents and flue gas heat recovery. The scenario using
 389 “new solvent” and heat recovery achieved 38%_{HHV} efficiency, which is 8% higher
 390 than the world average of 30% (LHV, or slightly <29%_{HHV}) for coal-fired power
 391 plants [111]. Thus, these efficiency enhancements address the energy penalty
 392 issues caused by co-firing biomass and solvent regeneration.

Figure 6: Efficiency ($\%_{\text{HHV}}$) of a 500 MW BECCS plant using different capture solvents and HR at 50% biomass co-firing, compared with average power plant efficiencies worldwide. NS = new solvent, HR = heat recovery, BAU = business as usual.



393 3.5. Carbon intensity

394 As the biomass co-firing proportion is increased, the BECCS plant be-
 395 comes more carbon negative (Figure 7). Additionally, the carbon negativity
 396 of the system increased as the plant efficiency decreased. As shown in Table
 397 4 at 50% biomass co-firing, the low efficiency MEA system captured -296 kg
 398 CO₂/MWh, whereas the ‘new solvent’ system with heat recovery captured -244
 399 kg CO₂/MWh. Although counter-intuitive, the rationale is that a system with
 400 higher efficiency would burn less fuel per MWh of electricity generated, thus,
 401 less CO₂ is captured from the atmosphere. This highlights the importance of
 402 identifying an appropriate performance matrix for the evaluation of BECCS.

403 It is important to note that this conclusion holds as long as the CO₂ emissions
 404 associated with biomass supply chain – or biomass carbon footprint (BCF) –
 405 do not offset the amount of CO₂ removed by the power plant. There is a
 406 maximum carbon footprint value, which is a function of the biomass carbon
 407 content, co-firing proportion and capture rate, beyond which, the facility is no

Figure 7: System carbon intensity as a function of the biomass co-firing proportion %.

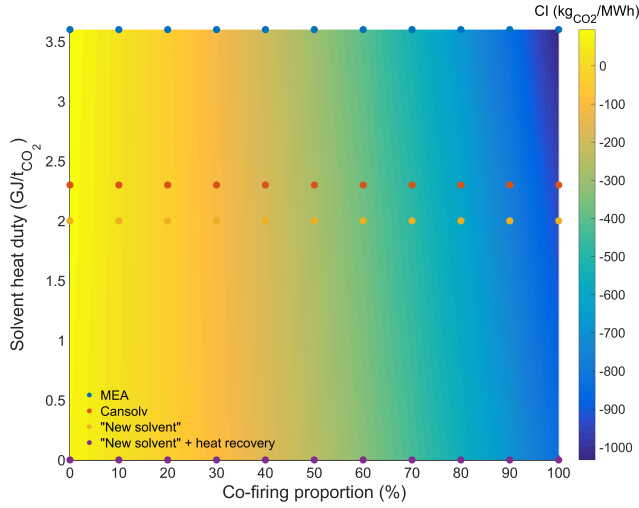


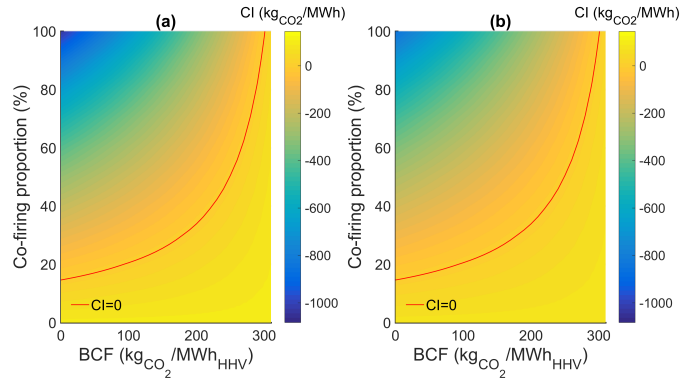
Table 4: Effect of solvent type on the carbon intensity and efficiency of a 500 MW BECCS plant co-firing 50% biomass with 90% CO₂ capture.

System	Efficiency (% _{HHV})	Carbon intensity (kg CO ₂ /MWh)
MEA	31.3	-296
Cansolv	33.7	-275
New solvent	34.3	-271
New solvent + heat recovery	38.0	-244

408 longer carbon negative. In this configuration, facilities that are more efficient
 409 at converting biomass into power will emit less CO₂ than their less efficient
 410 counterparts. The effect of biomass carbon footprint in kg_{CO₂}/MWh_{HHV} and
 411 co-firing proportion on the power plant carbon intensity is illustrated for wheat
 412 straw at 90% capture rate in figure 8a (low efficiency MEA process) and 8b
 413 (high efficiency new solvent process).

414 As can be observed in both figures, less efficient plants are slightly more car-
 415 bon negative for values below the BCF upper limit, and more carbon positive

Figure 8: System carbon intensity (CI) as a function of the biomass co-firing proportion % and biomass carbon footprint (BCF), for a MEA solvent (left) and a "new solvent" (right).



416 for values above the BCF upper limit compared to the more efficient plants.
 417 The BCF upper limit increases with co-firing, from 93 $\text{kg}_{\text{CO}_2}/\text{MWh}_{\text{HHV}}$ at 20%
 418 co-firing, to 301 $\text{kg}_{\text{CO}_2}/\text{MWh}_{\text{HHV}}$ at 100% co-firing. To put this in context,
 419 bioelectricity facilities are required to report emissions 60% lower than the EU
 420 mean electricity carbon intensity, according to the UK Bioenergy strategy [112].
 421 This translates into a limit of 285 $\text{kg}_{\text{CO}_2}/\text{MWh}_e$ for embedded GHG emission
 422 in the biomass supply chain, or a limit of 128 $\text{kg}_{\text{CO}_2}/\text{MWh}_{\text{HHV}}$ when assuming
 423 a maximal biomass conversion efficiency of 45%_{HHV}. These limits are over the
 424 BCF upper bound at 20% co-firing, which underlines the importance of consid-
 425 ering the biomass carbon footprint during identification of BECCS performance
 426 indicators.

427 4. BECCS performance matrix

428 The results are summarised into a BECCS performance matrix in Table 5,
 429 which illustrates the effect of key properties on the plant performance. The
 430 measures for performance in a BECCS system include: (i) pollutant reduction,
 431 (ii) energy efficiency, (iii) CO_2 negativity, or (iv) a combination of these.

Table 5: Performance matrix for a BECCS system, demonstrating the relationship between process properties and performance indicators. * For a power plant of a given capacity, the use of a fuel with lower HHV will require higher fuel firing rates, which will increase the flow rate of flue gas, leading to higher heat recovery.

Performance parameter	Ash content	Sulphur content	Moisture content	HHV	Biomass co-firing %	Solvent heat duty
NO _x ↓	↑	↓	↓	-	↑	-
SO _x ↓	↓	-	↓	-	↑	-
AFT ↑	↓	-	↓	↓*	↑	↓
Heat recovery ↑	↓	-	↓	↑	↓	↓
Efficiency ↑	↑	-	↑	↑	↑	↑
CO ₂ negativity ↑	-	-	↑	↓	↑	-
Exhaust gas flow rate ↑	-	-	↑	↓	↑	-

4.1. Reduction of pollutant emissions

The reduction of pollutant emissions from a power plant will be necessary for two possible reasons: (i) satisfy stringent emission regulations [43], or (ii) meet requirements of downstream processes for air pollution control [113, 114, 115]. Blending coal with biomass can be used as a means to reduce pollutant emissions of SO_X and NO_X [32], as the results from this study have demonstrated. Significant reductions to SO_X emissions can be achieved by increasing the biomass co-firing proportion (decrease in overall sulphur content). Fuel selection is another important factor; selecting coals that have low sulphur content can significantly reduce SO_X emissions. The SO_X levels ranged from 420 – 1180 ppm when high sulphur coal was co-fired with biomass, whereas the use of medium sulphur coal reduced SO_X emissions to <10 ppm (satisfying tolerance requirements of amines). Further reductions to SO_X can be achieved by utilising biomass with ash that contains alkali metal oxides (*e.g.*, CaO , MgO , Na_2O and K_2O). However, biomass ash containing silicates, chlorides and sulphates of potassium and calcium are undesirable as they increase slagging and fouling formation [42].

Although co-firing with fuels of low nitrogen content will contribute to reductions in NO_X emissions [103, 104], NO_X formation is mainly influenced by combustion conditions. To minimise NO_X emissions for each specific fuel blend, the combustion conditions require optimisation or reconfiguration of the burners. Some approaches used to reduce NO_X formation include fuel/air staging [110], or replacing N_2 gas with CO_2 (reduces NO_X formation from the N_2 in air) [41].

4.2. Improvement to power efficiency

To improve the efficiency of the BECCS systems and reduce the marginal cost of electricity generation, it is important to consider the following:

- Fuel selection for ideal combustion properties;
- High performance solvents (low heat duty and regeneration temperature);
- Heat recovery (increases with higher AFT and flue gas flow rate).

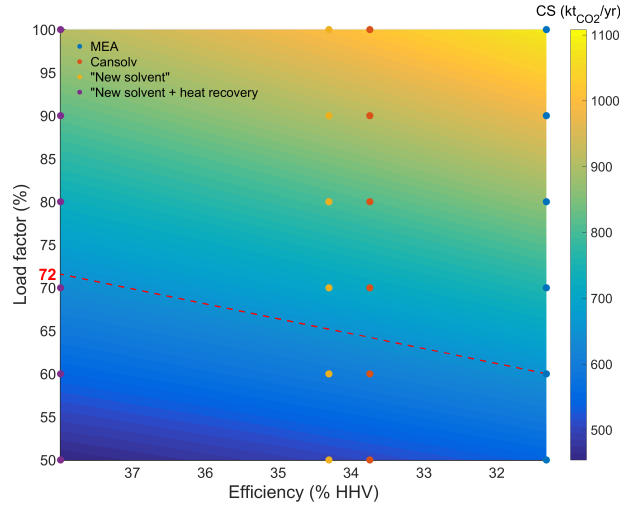
461 The fuel properties that would improve efficiency include higher HHV, lower
462 moisture and ash content. For instance, selecting coals with lower ash content,
463 whereas low moisture biomass with low tendency for ash deposition is ideal.
464 To minimise the energy penalty from CO₂ capture, utilise a high performance
465 solvent with low heat duty and low regeneration temperature.

466 Although the heat in the exhaust gas at the boiler exit is considered “waste”
467 from the perspective of electricity generation, this is useful energy, in terms of
468 quality and quantity, for solvent regeneration in the CO₂ capture process. The
469 implementation of a heat recovery system would require capital investment.
470 However, the advantage would be that heat recovery could allow BECCS to
471 operate without the efficiency penalty associated with CO₂ capture. Thus, en-
472 hancing heat recovery conditions can indirectly improve plant efficiency. How-
473 ever, some factors that improve heat recovery tend to reduce efficiency. For
474 instance, co-firing fuels with low HHV (*e.g.*, biomass) reduces plant efficiency,
475 however, requires higher fuel firing rate and increases the flue gas flow rate,
476 which in turn increases recoverable heat.

477 4.3. Enhancement of CO₂ negativity

478 The systems with lower efficiency will be more carbon negative (per MWh
479 basis) as they consume more biomass fuel, which captures and permanently
480 stores more CO₂ from the atmosphere. Consequently, enhancing the CO₂ nega-
481 tivity per MWh for a BECCS system would require implementing measures that
482 reduce efficiency. This may involve co-firing biomass with low fuel quality (*e.g.*,
483 high moisture and ash) or using the least efficient CO₂ capture system (*e.g.*, sol-
484 vents with high heat duty such as MEA). The use lower quality fuels can reduce
485 fuel costs. Also, the least efficient subcritical plants generally will have lower
486 capital costs compared to the high efficiency supercritical or ultrasupercritical
487 power plants [116], further cost reductions are possible by retrofitting the cur-
488 rent fleet of power plants for BECCS, most of which use subcritical technology
489 [117, 118]. However, it is essential to consider the impact of plant efficiency on
490 the annual dispatch load of the system.

Figure 9: Annual negative carbon emissions ($\text{kt}_{\text{CO}_2}/\text{yr}$), *i.e.*, avoided CO_2 emissions, as a function of the system efficiency ($\%_{\text{HHV}}$) and load factor ($\%$) of a BECCS system.



491 The load factor (annual capacity) of a power plant is a function of the plant
 492 efficiency. Systems that have higher efficiency will have lower marginal costs
 493 for electricity generation, which enables them to be economically competitive
 494 with other power generation technology. Therefore, enabling these plants to
 495 operate at higher load factors (higher annual dispatch factor) [27, 28]. Figure 9
 496 illustrates the influence of plant efficiency and annual capacity (load factor $\%$)
 497 on the annual avoided CO_2 emissions. The low efficiency MEA system captures
 498 $0.66 \text{ Mt}_{\text{CO}_2}$ annually at 60% capacity. However, the high efficiency system using
 499 new solvent and heat recovery capture the same amount of CO_2 annually but
 500 will operate at a capacity factor of 72%. Thus, increasing the CO_2 negativity
 501 (*e.g.*, decreasing plant efficiency) of a BECCS plant on a MWh basis will not
 502 necessary increase the annual CO_2 mitigation potential.

503 5. Conclusion

504 BECCS has the potential provide negative emissions, whilst providing reli-
 505 able firm power generation capacity. However, marginal costs are higher com-

506 pared to other power generation technologies due to the energy penalty incurred
507 from the CO₂ capture process and use of biomass fuel. This study evaluates
508 opportunities to improve the performance of BECCS through: (i) pollutant
509 reduction, (ii) efficiency enhancement, and (iii) CO₂ negativity.

510 Biomass co-firing with coal has been reported to provide significant reduc-
511 tions in SO_X and NO_X emissions. The thermochemical analysis demonstrates
512 that increasing biomass co-firing proportion reduced SO_X emissions due to the
513 decrease in fuel sulphur content. Fuel selection is also an important factor in
514 reducing SO_X formation. For instance, co-firing low sulphur coal reduced SO_X
515 by up to two orders of magnitude. Although some alkali metals are associated
516 with ash deposition problems, these ash component have a role in SO_X reduc-
517 tion. The NO_X emissions were mainly dependent on combustion conditions,
518 therefore, conditions may need to be calibrated to minimise NO_X formation.

519 Factors that enhance efficiency in a BECCS system include the use of high
520 performance solvents (low heat duty) and using heat recovery to supply energy
521 for solvent regeneration. The efficiency can be improved indirectly by increasing
522 the recoverable heat from the flue gas, *e.g.*, greater AFT or higher flue gas flow
523 rate. By using a high performance solvent with heat recovery, the BECCS
524 system could achieve an efficiency of 38%_{HHV}, higher than the current fleet of
525 coal-fired power plant with efficiencies ranging from 26 – 35%.

526 On the other hand, to achieve greater carbon negativity (on a per MWh
527 basis), a low efficiency system is more desirable due to increased consumption
528 of biomass fuel, which results in more CO₂ being captured and permanently
529 stored. Increasing the CO₂ negativity per MWh would involve using biomass
530 with high moisture and ash content or utilising the least efficient capture solvent,
531 *e.g.*, MEA. However, increasing CO₂ negativity on a MWh basis (*e.g.*, low
532 efficiency) will not equate to higher CO₂ mitigation. It is important to consider
533 the impact of efficiency on the dispatch rate of the system in the electricity
534 market. Typically, reducing the efficiency of a power plant will decrease its
535 annual capacity factor, lowering its dispatch rate in comparison to the high
536 efficiency systems.

537 The major barrier for commercial deployment of BECCS is the lack of eco-
538 nomic and political drivers, rather than technical (assuming fossil-fuel CCS is
539 now a proven technology) [26]. The potential role of BECCS in the electricity
540 grid is unique as it will provide reliable firm capacity, whilst also providing a
541 means to decarbonise the electricity sector [16, 17]. To enhance the economic
542 viability and encourage deployment of BECCS plant with higher CO₂ negativ-
543 ity, financial incentives for electricity generation with negative CO₂ emissions
544 would be necessary. To improve the commercial potential of BECCS, there is a
545 need to develop sustainable biomass supply chains and establish suitable CO₂
546 sequestration sites [26].

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